# Assessment of Riparian Ecosystem Integrity In the San Juan/San Mateo Watersheds, Orange County, California

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# Assessment of Riparian Ecosystem Integrity In the San Juan/San Mateo Watersheds Creek Watershed, Orange County, California

## 1.0 Special Area Management Plan

The Los Angeles District Corps of Engineers - Regulatory Branch is developing a Special Area Management Plan (SAMP) for the San Juan Watershed and portions of the San Mateo Watershed in Orange County, California. The Los Angeles District is conducting the SAMP in coordination with the existing and the proposed amendment to the Central - Coastal Natural Community Conservation Plan (NCCP).

The goal of the SAMP is to…"develop and implement a watershed-wide aquatic resource management plan and implementation program, which will include preservation, enhancement, and restoration of aquatic resources, while allowing reasonable and responsible economic development and activities within the watershed-wide study area" (Los Angeles District Corps of Engineers 1999). To achieve this goal, the aquatic resources within the San Juan/San Mateo Watersheds were identified, characterized, delineated (Lichvar 2000), and assessed at a planning level.

## 2.0 Project Objective

The overall objective of this project was to conduct a baseline assessment of riparian ecosystem integrity in the San Juan/San Mateo Watersheds under current conditions. Once completed, the information developed during the assessment will be used to evaluate the potential impacts of future development projects on riparian ecosystems in the watershed. A similar project has been completed for the San Diego Creek Watershed in Orange County (Smith 2000).

Three specific tasks were identified to meet the overall project objective. The first was to conduct a baseline assessment of riparian ecosystem integrity in the watersheds under current conditions. This was accomplished by dividing the riparian ecosystems into assessment units or "riparian reaches" and assessing each riparian reach using a suite of indicators of ecosystem

integrity (see the Assessment Indicators and Assessment Procedure sections below). Establishing baseline conditions allows for a comparison between riparian ecosystem assessment units under current conditions, and for a comparison of riparian ecosystem integrity under current and future conditions. Such comparisons will guide the decision-making process concerning future development projects by ensuring avoidance, minimization, and adequate mitigation of impacts to these resources, both individually and cumulatively.

The second task was to rank riparian reaches in terms of ecosystem integrity. Ranking was based on the ecosystem integrity indicator scores and hydrologic, water quality, and habitat integrity indices resulting from the baseline assessment. The rankings will be one of many factors used to evaluate various alternatives within the watershed.

The third task, which has not yet been completed, will be to determine which of several proposed alternative development scenarios would result in the least impact to riparian ecosystem integrity in the watershed. This will be accomplished by comparing the assessment indicator scores and integrity indices of riparian reaches under baseline conditions with the scores and indices of riparian reaches following the "simulation" of each proposed alternative scenario. Simulations will be based on an implementation of the changes that can be expected to occur in the context of each assessment indicator as a result of each proposed alternative development scenario.

## 3.0 Background, Definitions, and Assumptions

### 3.1 Riparian Ecosystems

Riparian ecosystems are linear corridors of variable width that occur along perennial, intermittent, and ephemeral streams (Williams 1978). Perennial streams are defined as streams in which base flow is maintained year round by groundwater. Intermittent streams are defined as streams in which ground water maintained base flow occurs intermittently at different times of the year. Ephemeral streams are defined as streams in which flow is attributable only to surface water runoff in response to precipitation. Two features that distinguish riparian ecosystems are the hydrologic interaction that occurs between the stream channel and adjacent areas through the periodic exchange of surface and ground water, and the distinctive geomorphic features and vegetation communities that develop in response to this hydrologic interaction (Richards 1982;

Harris 1987; Kovalchik and Chitwood 1990; Gregory et al. 1991; Malanson 1993; and Goodwin et al. 1997).

The hydrologic interaction between streams and adjacent riparian areas typically results in two distinct zones, although either zone may be narrow and seemingly absent under certain geologic or geomorphic conditions. The first zone, the active floodplain, includes the areas that are inundated by overbank flooding at least once every five years. This zone exhibits the fluvial features associated with recurring flooding such as point bars, areas of scour, sediment accumulation, natural levees, and debris wrack, and vegetation communities that are either short lived or able to survive the effects of frequent flooding (Figure 1).

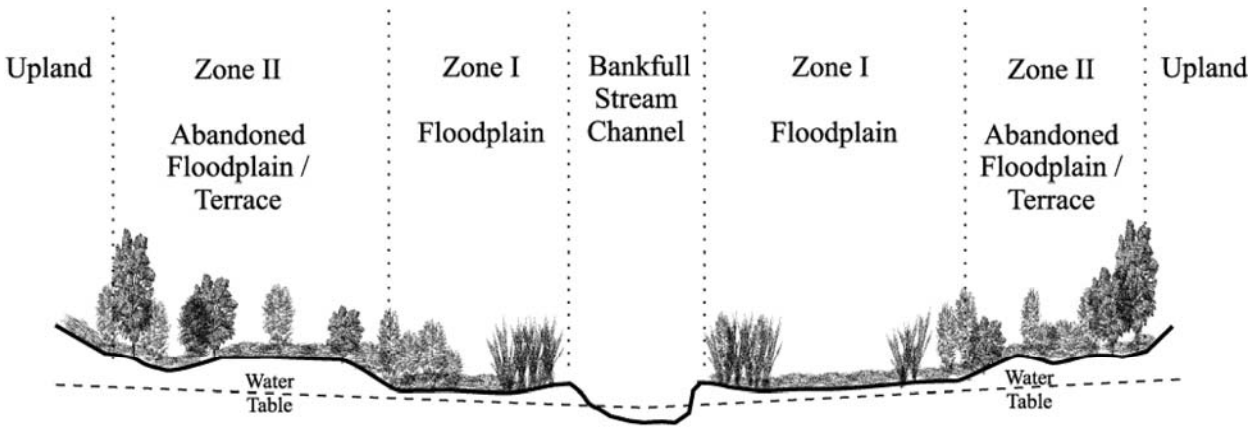


Figure 1. Generalized cross section of a riparian ecosystem

The second zone consists of abandoned floodplains and historical terraces formed by fluvial processes that took place under a historical, and often different, climatic condition or hydrologic regime (Knox et al. 1975; Graf et al. 1991; Rumsby and Macklin 1994). Under current climatic conditions and hydrologic regimes, these areas are flooded less frequent, during larger magnitude events (Dunn and Leopold 1978). Vegetation communities in this zone are generally composed of woody perennials that require the higher water table present in the riparian zone, and are capable of surviving, or reestablishment, after floods. For the purposes of this project, riparian ecosystems were defined from a functional perspective as the areas along perennial, intermittent, and ephemeral streams where the interaction with surface and groundwater results in distinctive geomorphic features and vegetation communities. Under natural circumstances, the riparian ecosystem includes the bank full stream channel, active floodplain, and less frequently flooded, historical floodplains/terraces.

### 3.2 Waters of the United States Versus Functional Riparian Ecosystems

Waters of the United States (WoUS) are the areas subject to regulation under Section 404 of the Clean Water Act (33 CFR Part 328.3). Wetlands are a subset of WoUS, and throughout this discussion, the term WoUS should be interpreted as including wetlands. Two categories of WoUS occur in association with southern California riparian ecosystems. The first category, non-wetland waters, are the areas along perennial, intermittent, and ephemeral stream channels that exhibit a distinct bed and bank, but fail to meet one or more of the hydrologic, hydrophytic vegetation, and hydric soils criteria outlined in the Corps of Engineers Wetlands Delineation Manual (Environmental Laboratory 1987). The second category of wetlands, are the areas that meet all of the hydrologic, hydrophytic vegetation, and hydric soils criteria

It is important to note the "functional" riparian ecosystem, as defined for this project, has no special recognition, meaning, or status in the context of the 404 Program. While functional riparian ecosystems normally include all WoUS regulated under the 404 Program and California Department of Fish and Game (CDFG) 1600 Program, the riparian ecosystem at times includes areas that fall outside the jurisdiction of one or both of these programs. Consequently, there is not necessarily a one-to-one spatial correspondence between riparian ecosystems and WoUS in the watershed. This lack of spatial correspondence is common in the arid southwestern United States where the active floodplain or historical terrace portion of the riparian ecosystem often meets one or two of the delineation criteria, but fails to meet all three delineation criteria necessary to qualify as a regulated wetland.

The spatial inconsistency between WoUS and riparian ecosystems is a result of the relatively generic hydrologic, hydrophytic vegetation, and hydric soil delineation criteria developed under the 404 Program for use in the wide variety of wetland types that occur in the United States. These generic delineation criteria necessarily ignore the unique way in which specific characteristics and processes contribute to the creation and maintenance of riparian as well as other functional wetland ecosystems. The intra- and inter-regional insensitivity of the generic delineation criteria is widely recognized, and while the need for regionalization of delineation criteria has been identified (Committee on Characterization of Wetlands 1995), no solution to this formidable task has been developed, much less implemented.

The spatial inconsistency is problematic in the context of the mandate to assess functions of WoUS as part of the 404 permit review process. Clearly, an assessment cannot be accomplished by considering only the characteristics and processes of WoUS proper. This is because the functions of WoUS are influenced by the characteristics and processes taking place in the riparian ecosystem, as well as the upland areas adjacent to the riparian ecosystem, and the drainage basin of the riparian ecosystem (Kratz et al. 1991; Hornbeck and Swank 1992; Bedford 1996).

A solution for meeting this challenge was outlined as part of the Hydrogeomorphic (HGM) Approach (Smith et al. 1995). In the HGM Approach, characteristics and process occurring in the functional ecosystem, as well as the adjacent upland areas and the drainage basin are considered during the assessment process. However, when applying the results of the assessment in the context of the 404 permit review process, the results are applied only WoUS. This project used a similar approach in that the influence of the riparian ecosystem, adjacent uplands, and drainage basin were considered in assessing riparian ecosystem integrity. Consequently, when applying the results of the assessment, consistency with policies and assumptions of the SAMP, the 404 permit review process, Section 7 consultation, or the California Department of Fish and Game 1600 Program must be taken into account.

### 3.3 The Riparian Reach Assessment Unit

Due to the large size of the project watershed (>450 $km^2$), inherent variability of riparian ecosystems, and differential nature of historical impacts to riparian ecosystems in the watershed, the selection of an appropriate spatial unit for assessing riparian ecosystems was critical. The assessment unit selected was called the riparian reach (RR), and defined as a segment of the main stem, bank full stream channel and adjacent riparian ecosystem that was relatively homogenous with respect to geology, geomorphology, channel morphology, substrate type, vegetation communities, and cultural alteration (Olson and Harris 1997) (Figure 2).

In association with each riparian reach two other areas were defined. These included the "local drainage" of the riparian reach (RRLD) and the "drainage basin" of the riparian reach (RRDB). The local drainage of a riparian reach included the area from which surface water drained directly to the main stem channel or tributaries that entered the main stem channel in the

riparian reach. The local drainage did not include areas that drained to the main stem channel associated with upstream riparian reaches (Figure 2). The drainage basin of a riparian reach included the local drainage of the riparian reach in addition to the local drainage area of all upstream riparian reaches (Figure 2). The local drainage and the drainage basin provide the spatial context that is used for several "indicators" or riparian ecosystem integrity (see Sections 4.0 and 5.0).

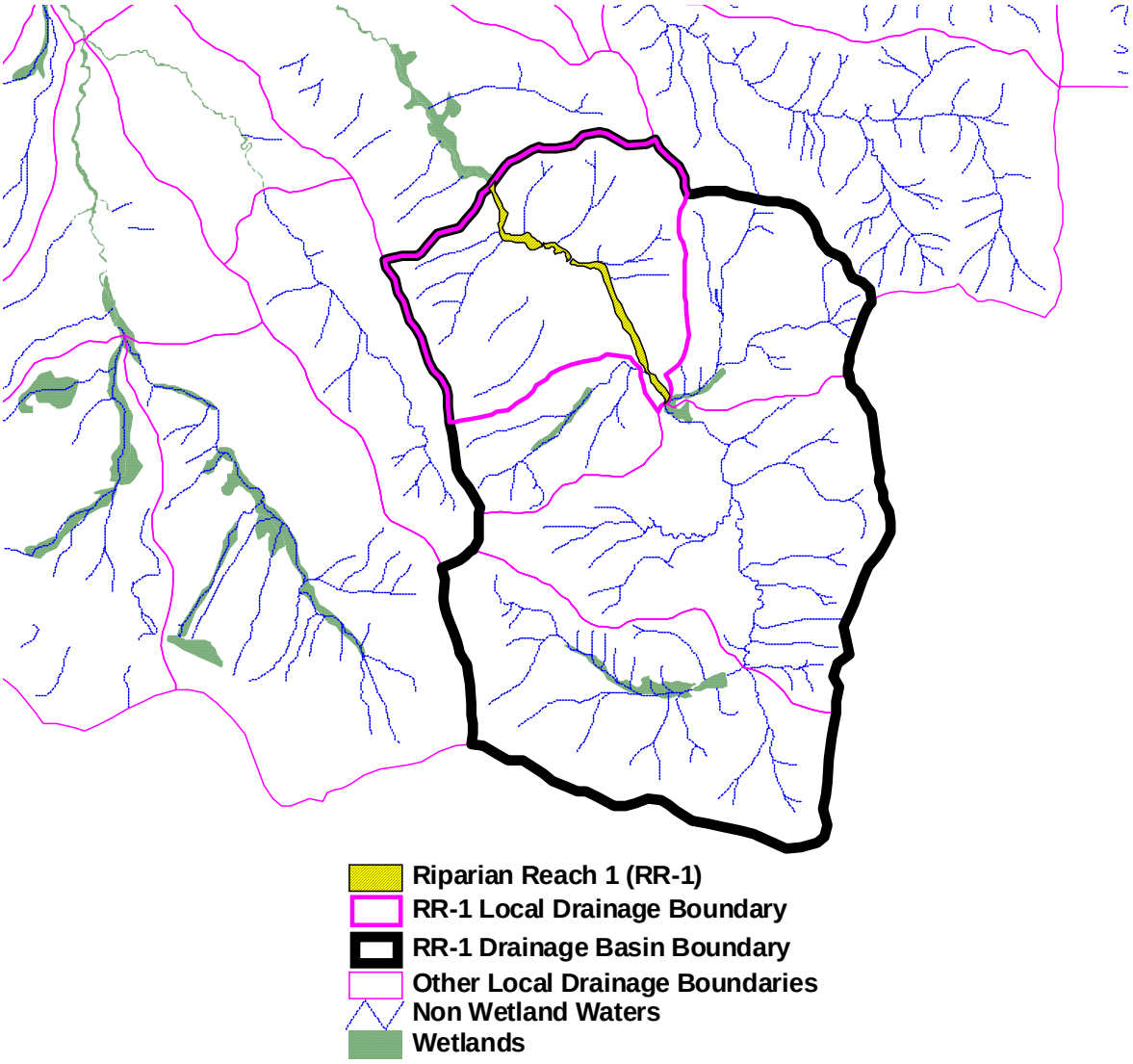


Figure 2. Illustration of riparian reach and its associated local drainage and drainage basin

### 3.4 Riparian Ecosystem Integrity and Assessment Endpoints

Much has been written in the last few years about the concepts of ecological or ecosystem health and integrity (Rapport 1989; Costanza, Norton and Haskell 1991; Suter 1993; Scrimgeour and Wicklum 1996; Karr 1999). The two terms are often used interchangeably; however, the distinction made by Karr (1996) is instructive, and important in interpreting and applying the mandate of the Clean Water Act. Health refers to a flourishing condition, well-being, or vitality (Guralnik and Friend 1968). Integrity, on the other hand, refers to the quality, or state of being complete, and implies correspondence with a natural or original condition. Based on these distinctions, a cornfield, pine plantation, commercial nursery, and other culturally altered ecosystems qualify as healthy, but do not qualify as ecosystems with high integrity. For this project, riparian ecosystems with high integrity were defined as riparian ecosystems that support a balanced, diverse, and adaptive biological community resulting from natural evolutionary and biogeographic processes, and exhibit the full range of physical, chemical, and biological attributes and processes that characterized riparian ecosystems in the region, over short and long term cycles, prior to cultural alteration.

The concept of ecosystem integrity is difficult to define because of its abstract nature. However, it is even more difficult to assess because of the wide variety of characteristics and processes that influence riparian ecosystem across multiple spatial scales (i.e., riparian reach, local drainage, and drainage basin). Consequently no single, direct measure of ecosystem integrity exists. This project focused on three quantities of interest, or assessment "endpoints," to represent riparian ecosystem integrity (Liebowitz and Hyman 1999). These included hydrologic, water quality, and habitat integrity. The selection of these endpoints followed directly from the mandate in Section 101(a) of the Clean Water Act to "…restore and maintain the chemical, physical, and biological integrity of the Nation's waters". Each of these assessment endpoints is defined and discussed in greater detail in the Assessment Endpoint section below.

### 3.5 Use of Indicators to Assess Hydrologic, Water Quality, and Habitat Integrity

Once assessment endpoints have been selected, metrics must be selected to assess the endpoints. There are two categories of metrics. "Direct metrics" are a quantitative direct

measure of an endpoint. For example, cubic feet per second, is a direct metric that measures the stream discharge endpoint. Direct metrics can be identified when the assessment endpoint (i.e., attribute or process) is a narrowly defined and directly measurable. Direct metrics cannot be identified for broadly defined, abstract concepts like ecosystem integrity.

The second category of metric is the "indirect metric" or "indicator". Indicators are qualitative or quantitative measures that are in some way related to an assessment endpoint. Indicators are used to assess complex, or abstract, endpoints for which no direct metric exists. However, indicators are also frequently used when direct measures are too difficult or costly. For example, tree basal area is a metric used to indicate the endpoint of tree biomass. Many existing biological/ecological assessment methods use indicators for these reasons. For example, the Habitat Evaluation Procedure (HEP) (USFWS 1980) has used habitat characteristics as indicators for more than 25 years to assess a "habitat suitability" endpoint in lieu of the more difficult and time consuming task of sampling animal populations directly (USFWS 1980). Indicators are used in a similar fashion in the Index of Biological Integrity (IBI) and related methods (Karr and Chu 1997), the instream flow incremental methodology (IFIM) (Bovee 1986), the Synoptic Approach (Leibowitz et al. 1992; Abbruzzese and Leibowitz 1997), and the Hydrogeomorphic (HGM) Approach (Smith et al. 1995).

Liebowitz and Hyman (1999) make an important distinction between "confirmed" and "judgment" indicators. Confirmed indicators are those in which the relationship between the indicator and endpoint can be described in a precise manner (i.e., mathematical) with a known level of statistical confidence. Judgment indicators, on the other hand, are those in which the relationship between the indicator and endpoint is less precisely defined. The relationship is typically based on trends or patterns from the literature, field observations, or professional judgment. Given adequate research, the assumed relationship between a judgment indicator and an endpoint can be confirmed. For example, it is possible to define a relatively precise mathematical relationship between land use and water quality in a watershed using numerical modeling methods (Hamlett et al. 1992). The use of confirmed versus judgment indicators represents a tradeoff in terms of the degree of certainty of the relationship between the indicator and endpoint, and the ability to obtain the information necessary to assess selected endpoints. Some authors have questioned the use of judgment indicators (Conroy and Noon 1996,

Schumaker 1996). However, in real world situations the use of judgment indicators is often unavoidable given time and resource constraints, the lack of existing confirmed indictors, or the unavailability of quantitative data necessary to develop a confirmed indicator (Abbruzzese and Leibowitz 1997). Each of the indicators selected to assess hydrologic, water quality, and habitat integrity are discussed in greater detail in the Assessment Indicators Section below.

### 3.6 Reference Condition

In order to assess the integrity of a riparian reach, a standard of comparison or "reference condition" must be defined. The reference condition serves two purposes. First, it provides a concrete or virtual representation of the conditions, across multiple spatial scales, under which riparian ecosystems achieve and sustain a high level of integrity. Second, the reference condition provides a starting point from which to define and scale the relationship between the indicators and assessment endpoints.

Several different reference condition scenarios were suggested and considered for this project. Two of these were the "culturally unaltered" and "least culturally altered" reference condition. In southern California riparian ecosystems, the culturally unaltered reference condition implies conditions that existed prior to grazing, agriculture, fire suppression, water resource management, transportation corridors, urbanization, and other cultural alterations that can be identified. It is synonymous with what McCann (1999) referred to as pre-Columbian, meaning the conditions that existed prior to the influence of European explorers and subsequent immigrants. The least culturally altered reference condition refers to those conditions that currently exist in a watershed or region and most closely reflect culturally unaltered conditions.

Culturally unaltered was selected as reference condition for this project for several reasons. First, it represents the physical, chemical, and biological conditions under which riparian ecosystems have evolved naturally, and therefore, presumably represents the physical, chemical, and biological conditions that the Clean Water Act mandates should be maintained. While it can be argued that the culturally unaltered reference condition does not exist in southern California due to widespread existence of grazing, fire suppression, urban development, non-point air pollution, the disruption of historical metapopulation dynamics (Hastings and Harrison 1994), and a host of other factors, it is possible to make reasonable speculations as to what

culturally unaltered conditions were like (Sedell and Luchessa 1981; Schubauer-Berigan 2000). It can also be argued while it is impossible to restore culturally unaltered conditions, it may be feasible to restore some of the larger, isolated and remote areas to a condition that functionally approximates the culturally unaltered condition given adequate time and resources, and appropriate management.

In the restoration context, a reference condition based on the culturally unaltered scenario provides an appropriate target for restoring ecosystem integrity and function. On the other hand, a restoration target based on the least culturally altered reference condition provides an entirely arbitrary, and often inappropriate target with the potential to "successfully" restore riparian ecosystems with low ecosystem integrity and function, and no natural corollary.

The second reason for selecting culturally unaltered as the reference condition is the generally unappreciated advantage, both in terms of interpretation and comparability of results, of using the "absolute" standard of comparison culturally unaltered represents versus the "relative" standard of comparison least culturally altered represents. To illustrate this advantage, consider the following scenario. Assessments of ecosystem integrity are done on riparian ecosystems in two watersheds, one heavily urbanized and the other occupying a wilderness area without roads. Two assessments are done in each watershed. The first assessment uses culturally unaltered conditions as the reference condition, and second uses least culturally altered conditions as the reference condition. Indices of ecosystem integrity are generated for both assessments ranging from 1 to 10 with an index of 1 indicating low integrity. In the first assessment, using culturally unaltered conditions as the reference condition, the indices for the urban watershed are likely to be at the lower end of the index range, while the indices for the wilderness watershed are likely to be in the higher end of the index range. These results are intuitively reasonable, and in reality correct, because heavily urbanized watersheds have significantly less ecosystem integrity than wilderness area watersheds due to changes in land use, stream channelization, loss of habitat, and other factors.

Now consider the results for the second assessment using least culturally altered conditions as the reference condition. Indices for the urban watershed will be at the high end of the index range, because least culturally altered conditions, specific to the urban watershed, were used to scale indicators of ecosystem integrity. Indices for the wilderness watershed will be also be at

the high end of the index range for the same reason. However, these results are not intuitive because, using the foregoing definition of ecosystem integrity, the urban watershed in reality has a lower level of ecosystem integrity than the wilderness area, despite the fact that the indices of ecosystem integrity indicate there is little difference between the two. The non-intuitive nature of these results, and the inability to compare areas makes the use of the relative, least culturally altered reference condition, problematic at best.

The third reason for selecting culturally unaltered as the reference condition was the ability to define a culturally unaltered condition for the indicators of riparian ecosystem integrity without extensive reconnaissance in the watershed prior to conducting the assessment. For example, in the case of the indicators related to land use, it was reasonable to assume that under the culturally unaltered condition no grazing, agriculture, transportation, or urban development land uses existed. Similarly, in the case of the altered hydrologic conveyance indicator, it was reasonable to assume that under culturally unaltered conditions, stream channels were not straightened, lined, impounded, or underground. The same could not be said for defining the least culturally altered condition. In order to define least culturally altered condition for assessment indicators it would be necessary to conduct reconnaissance in the watershed, prior to conducting the assessment, to determine the range of cultural alteration that existed and what represented least culturally altered condition.

## 4.0 Assessment Endpoints

The assessment endpoints selected to represent riparian ecosystem integrity for this project were hydrologic integrity, water quality integrity, and habitat integrity. Each of these endpoints in defined in the following sections and discussed in terms of assessment indicators. The indicators are defined and discussed in greater detail in the Assessment Indicators Section below.

### 4.1 Hydrologic Integrity

Riparian ecosystems with high hydrologic integrity exhibit the range of frequency, magnitude, and temporal distribution of stream discharge, and surface and subsurface interaction between the stream channel, floodplain, and terraces, that historically characterized riparian ecosystems in the region (Bedford 1996, Poff et al. 1997, Richter et al. 1997). In the arid southwest, this translates into seasonal intermittent, ephemeral, or low flow periods, with annual

bankfull discharges superimposed on a background of episodic, and often catastrophic, larger magnitude floods that inundate historical terraces (Graf 1979; Graf 1988; Harris 1987; Fisher et al. 1982; Friedman et al. 1996a, Friedman et al. 1996b).

In selecting indicators to assess hydrologic integrity, two groups of characteristics and processes were considered. The first group focused on the factors that influence frequency, magnitude, and temporal distribution of stream discharge, and the second group focused on the factors that influenced the hydrologic interaction between the stream channel, floodplain, and historical terraces. Direct measures of stream discharge are unavailable at the riparian reach scale in these watersheds. Consequently, several indicators were selected at the drainage basin scale with the assumption that an indirect estimate of deviation from reference condition can be made based on changes in specific characteristic and processes of a drainage basin such as interception, infiltration, evapotranspiration, percolation, groundwater flow, and surface water flow overland and in channels. Cultural alteration of the drainage basin alters these characteristics and processes and consequently stream discharge. While it is difficult to quantify the exact nature of the relationship between specific drainage basin characteristics, as represented by the indicators, and stream discharge, it can generally be shown that as cultural alteration of a watershed increases, so does the deviation from short and long-term historical patterns of frequency, magnitude, and distribution of stream discharge.

Four indicators of hydrologic integrity were selected to reflect degree of cultural alteration in a drainage basin with the potential to influence stream discharge included:

- Altered Hydraulic Conveyance (RRDB Scale)
- Surface Water Retention (RRDB Scale)
- Perennialized Stream Flow (RRDB Scale)
- Import, Export, or Diversion of Surface Water (RRDB Scale)

A frequency, magnitude, and distribution of stream discharge in a riparian reach that is similar to the historical range of conditions does not alone, ensure hydrologic integrity. Hydrologic integrity also depends on maintaining the interaction between the stream channel, floodplain, and terraces of the riparian ecosystems through overbank and subsurface flows. This interaction is critical to the maintenance of riparian plant communities, sediment storage, carbon

dynamics, biogeochemical processes, and other characteristics and processes of riparian ecosystems.

Two indicators were selected to represent the degree of interaction between the stream channel and the floodplain included:

- Altered Hydraulic Conveyance (RR Scale)
- Floodplain Interaction (RR Scale)

These six indicators were used to calculate the Hydrologic Integrity Index for a riparian reach using the procedure described Section 6.4: Assigning Indicator Scores and Calculation of Indices.

### 4.2 Water Quality Integrity

Water quality integrity was defined as exhibiting a range of loading in the pollutant categories of nutrients, pesticides, hydrocarbons, and sediments that are similar to those that historically characterized riparian ecosystems in the region. Assessing changes in the range of loading in each pollutant category can be determined directly by comparing data on current loading with data on historical loading when such data is available. While there is some historical and recent monitoring data available for a limited number of stations in the watershed, little or no loading data is available at the riparian reach scale. Consequently, the assessment of water quality integrity was based on indicators of drainage basin and riparian reach characteristics that have been shown to influence water quality integrity.

Three groups of factors were considered in selecting indicators for the water quality integrity endpoint. The first group focus on whether or not the changes in land use in the drainage basin had the potential to increase sources of pollution compared to the reference condition. The second group focused on whether or not the stream channel delivery system had changed in relation to reference condition in terms of frequency, magnitude, and temporal distribution of stream flow (Kuenzler 1977). The third group focused on whether or not changes in land use in the areas adjacent to the stream, or the loss of a hydrologic connection between the stream channel and the floodplain had decreased the likelihood of pollutants being physically captured or biogeochemically processed compared to reference condition. A number of studies have shown that cultural alteration of these factors can lead to increased loading in one or more

pollutant categories (Osborne and Wiley 1988; Allan and Flecker 1993; Hunsaker and Levine 1995; Perry and Vanderklein 1996; Richards et al. 1996; Allen et al. 1997; Bolstad and Swank 1997; Johnson et al. 1997; Wang et al. 1997; Miltner and Rankin 1998; Trimble 1997; Basnyat et al. 1999).

One indicator of water quality was selected to reflect the condition of in land use in the drainage basin.

- Land Use / Land Cover (RRDB Scale)

This indicator was based on averaging the indicator valued for four sub-indicators including:

- Land Use / Land Cover – Nutrient Increase (RRDB Scale)
- Land Use / Land Cover – Pesticide Increase (RRDB Scale)
- Land Use / Land Cover – Hydrocarbon Increase (RRDB Scale)
- Land Use / Land Cover – Sediment Increase (RRDB Scale)

Five indicators were selected to reflect the condition of the stream system that transports pollutants. They are the same indicators used to assess hydrologic integrity with the exception of Floodplain Interaction and included:

- Altered Hydraulic Conveyance (RRDB Scale)
- Altered Hydraulic Conveyance (RR Scale)
- Surface Water Retention (RRDB Scale)
- Perennialized Stream Flow (RRDB Scale)
- Import, Export, or Diversion of Surface Water (RRDB Scale)

Three indicators of water quality were selected to reflect the condition of riparian ecosystem with respect to it ability to physically capture and biogeochemically process pollutants. They include:

- Floodplain Interaction (RR Scale)
- Sediment Regime (RR Scale)
- Area of Native Riparian Vegetation (RR Scale)

These nine indicators were used to calculate the Water Quality Integrity Index for a riparian reach using the procedure described in Section 6.4: Assigning Indicator Scores and Calculation of Indices.

### 4.3 Habitat Integrity

Riparian ecosystems with habitat integrity exhibit the quality and quantity of habitat necessary to support and maintain a balanced, integrated, adaptive biological system having the full range of characteristics, processes, and organisms at the site specific, landscape, and watershed scales that historically characterized riparian ecosystems in the region. Several factors were considered in selecting indicators of habitat integrity including the spatial extent and quality of riparian habitat, the "connectedness" of riparian habitats at the riparian reach and drainage basin scales, and the spatial extent and quality of upland habitat in the landscape adjacent to riparian ecosystems.

- Area of Native Riparian Vegetation (RR Scale)
- Riparian Corridor Continuity (RR Scale)
- Riparian Corridor Continuity (RRDB Scale)
- Land Use / Land Cover at Riparian Ecosystem Boundary (RRLD Scale)
- Land Use / Land Cover in 100m Buffer Around the Riparian Ecosystem (RRDB Scale)

These five indicators were used to calculate the Habitat Integrity Index for a riparian reach using the procedure described in Section 6.4: Assigning Indicator Scores and Calculation of Indices.

## 5.0 Assessment Indicators

The selection of assessment indicators was based on the identification of important characteristics and processes that influence assessment endpoints. Potential indicators were gleaned from a review of existing assessment methods (Dinius 1987; Lee et al. 1997; Ladson et al. 1999). Further investigation of the literature on riparian ecosystems, field observations, and the collective experience of individuals participating in the project provided additional potential indicators. In selecting indicators, the objective was to directly or indirectly capture to the greatest degree possible the full range of characteristics and processes that influence hydrologic,

water quality, and habitat integrity of riparian ecosystems at the riparian ecosystem, adjacent upland, and drainage basin.

Several other factors influenced the final selection process. First, was the need to match the project objectives of establishing baseline conditions and the ability to make comparisons between riparian ecosystems with available data, time and resources. Other factors included the large project area (>450 $km^2$), a short time frame and limited budget, the lack of quantitative data at the riparian reach assessment unit scale, and the lack of existing confirmed indicators. Another factor was the requirement to develop an open and easily understood approach that would allow participation and input from multiple stakeholders representing a range of perspectives from the development community to federal agencies charged with the protection of sensitive, threatened, and endangered species. Ultimately, a balancing of all these factors led to the selection of the indicators described below.

Each of the indicators are defined in the following sections and discussed in terms of the relationship between the indicator and relevant assessment endpoints. In addition, the method used to measure the indicator and assign an indicator value is described, along with the reference condition and range of indicator values used to assign indicator scores. Throughout this section the subscripts following the indicator acronym designates the spatial scale at which the indicator is applied with "RR" indicating the riparian reach scale, "RRLD" indicating the local drainage spatial scale, and "RRDB" indicating the drainage basin spatial scale.

### 5.1 Altered Hydraulic Conveyance ($AHC_{RR}$ / $AHC_{RRDB}$)

Altered Hydraulic Conveyance indicates the degree to which engineering techniques have been used to "improve" the capacity of channels in a riparian reach or drainage basin to convey surface water downstream. The engineering techniques involve reducing the frictional resistance (i.e., roughness) caused by channel substrate, vegetation, woody debris, and other objects in the channel (Barnes 1967), minimizing the wetted perimeter, and/or shortening the length of a channel. Specific techniques include dredging, straightening, hardening, and lining of the stream channel as well as the removal of vegetation (Galay 1983, Brookes 1988).

Increasing the volume of water and velocity at which water is conveyed downstream can result in a significant change in the hydrologic regime, and hence hydrologic integrity, in the

riparian reach where the alteration occurs as well as in upstream and downstream reaches. For example, removal of vegetation decreases channel stability and increases erosion by reducing the resistance afforded by the network of plant roots, and by increasing the velocity and consequently the erosive force of water in the channel. A straightened stream reach will typically respond by incising to reestablish a more energy efficient and stable channel slope (Shankman and Samson 1991). This in turn initiates headcutting and increased erosion upstream. Downstream of an altered stream channel the hydrologic regime can also be affected in terms of increased peak discharges, a decrease in channel stability, and an increase in erosion due to increased water velocity.

This indicator was measure as the percent of the main stem channel through the riparian reach with altered hydraulic conveyance. At the riparian reach scale, aerial photography and field observations were used to estimate the value of the metric. At the drainage basin scale, the indicator was calculated as the weighted average of the percent of altered hydraulic conveyance for all riparian reaches in the drainage basin of the riparian reach. In other words,

$$\sum_{i=1}^{i_n}\left[\frac{AHC_{RR} * \left(ML_{RR} / ML_{DB}\right)}{100}\right] \qquad (1)$$

Where: $AHC_{RR}$ = % of main stem in a riparian reach with altered hydraulic conveyance

$ML_{RR}$ = Length of main stem channel in a riparian reach

$ML_{DB}$ = Length of main stem channel of all riparian reaches in drainage basin

The reference condition was defined as <5% of the main stem channel in riparian reach with altered hydraulic conveyance. Indicator scores were assigned based on range of indicator values in Table 1.

Table 1. Range of indicator values for scaling the altered hydraulic conveyance indicator

| Indicator Value Range | Score |
|---|---|
| <5% of riparian reach main stem/drainage basin with AHC | 5 |
| >5 and <15% of riparian reach main stem/drainage basin with AHC | 4 |
| >15 and <30% of riparian reach main stem/drainage basin with AHC | 3 |
| >30 and <50% of riparian reach main stem/drainage basin with AHC | 2 |
| >50% of riparian reach main stem/drainage basin with AHC | 1 |

### 5.2 Surface Water Retention ($SWR_{RRDB}$)

Surface Water Retention indicates the degree to which the hydrologic regime in a riparian reach has been altered as a result of short and long-term storage of surface water in reservoirs, lakes, sediment basins, retention ponds, and similar surface water storage facilities. Streams in arid regions are disturbance-dominated systems (Resh et al. 1988; Power et al. 1988, 1996; Rood and Mahoney 1990). During flash floods, stream discharge can increase by several orders of magnitude causing aquatic organism mortality, destruction of riparian vegetation, and changes in channel morphology. The biological components of riparian ecosystems have adapted to these episodic cycles of disturbance, and developed a variety of mechanisms that make it possible to survive and indeed flourish where other organisms cannot. Short and long-term retention of surface water in storage facilities can significantly alter the characteristic pattern of discharge over the water year (Cushman 1985; Bain et al. 1988; Dynesius and Nilsson 1994; Ligon et al. 1995; Poff et al. 1997; Hadley and Emmett 1998). Most importantly, it eliminates the low frequency, high volume discharges that reset the system (Hawkins et al. 1997). However, it can also lead to perennialization of streamflow, change the pattern of seed distribution, germination, and survival, and change a variety of other physical and biological processes necessary to perpetuate the riparian ecosystem (Hynes 1975; Warren 1979; Lotspeich and Platts 1982; Frissell et al.1986; Kondolf et al.1987; Debano and Schmidt 1989; Stromberg and Patton 1991; Johnson 1994; Power et al. 1996; Kershner 1997; Kondolf 1997; Richter et al. 1997).

This indicator was measured as the percent of the drainage basin of a riparian reach upstream of a lake, reservoir, dry dam, sediment basin, retention pond, or similar facility capable of storing surface water from several days to months. The total area within each drainage basin upstream of the downstream extent of all storage facilities was determined using the ArcView GIS themes of riparian reaches, surface water retention facilities, and USGS 7.5 minute topographic map. Using the theme of surface water retention structures and a topographic map as background, the reach theme was split along topographic boundaries at the downstream extent of the retention structures. Upstream areas above these reach segments were calculated and summed across the drainage basin to determine the metric value.

The reference condition was defined as <5% of the drainage basin of a riparian reach upstream of a lake, reservoir, dry dam, sediment basin, retention pond, or similar facility capable

of storing surface water from several days to months. Indicator scores were assigned based on the range of indicator values in Table 2.

Table 2. Ranges of indicator values for scaling the surface water retention indicator

| Indicator Value Range | Score |
|---|---|
| <5% of drainage basin drains to surface water storage facilities | 5 |
| >5 and <15% of drainage basin drains to surface water storage facilities | 4 |
| >15 and <30% of drainage basin drains to surface water storage facilities | 3 |
| >30 and <50% of drainage basin drains to surface water storage facilities | 2 |
| >50% of drainage basin drains to surface water storage facilities | 1 |

### 5.3 Perennialized Stream Flow ($PSF_{RR}$ / $PSF_{RRDB}$)

Perennialized Stream Flow indicates the degree to which the hydrologic regime of a riparian reach has been altered by a supplementary supply of surface water resulting from cultural activites such as irrigation. Perennialization refers to the conversion of intermittent or ephemeral stream channels to a perennial stream through the addition of surface water flow (usually at low levels) in a stream channel from artificial supplies of surface water. The supply of water usually occurs in the form of irrigation or treated return water. In arid regions, perennialization facilitates a shift in plant and animal community composition away from what normally occurs in a riparian reach that is not perennialized. Perennialization also has the potential to affect physical and chemical processes in riparian ecosystems.

This indicator was measured as the percent of the main stem channel through a riparian reach that exhibited perennialized stream flow due to supplementary sources of water at the time of the field visits, or showed evidence of perennialized stream flow (i.e., occurrence of *Typha* sp., *Carex* sp. and/or other emergent aquatic species). Field observations and aerial photographs were used to assign a value to the indicator. The evidence used to identify a stream as perennialized was the presence of low flow during dry periods. Other types of evidence included nutrient enrichment based on the presence of blue-green algae and vascular species such as *Typha* sp., outfall pipes and other inlet structures entering a reach, residential developments and golf courses in the drainage basin, interbasin transfer import points, and the lack of evidence of a natural source of low flow.

The reference condition was defined as <5% of the main stem channel of a riparian reach with perennialized stream flow. Indicator scores were assigned based on the range of indicator values in Table 3.

Table 3. Range of indicator values for scaling the perennialized stream flow indicator

| Indicator Value Range | Score |
|---|---|
| <5% of stream channel exhibiting perennialized flow | 5 |
| >5 and <15% of stream channel exhibiting perennialized flow | 4 |
| >15 and <30% of stream channel exhibiting perennialized flow | 3 |
| >30 and <50% of stream channel exhibiting perennialized flow | 2 |
| >50% of stream channel exhibiting perennialized flow | 1 |

### 5.4 Import, Export, or Diversion of Surface Water ($IED_{RRDB}$)

Import, Export, or Diversion of Surface Water indicates the degree to which the hydrologic regime of a riparian reach has been altered as a result of import, export, or diversion of surface water. Inter-basin import and export of surface water, and the intra-basin diversion of water for public water supply, irrigation, and ground water recharge is common in the arid western United States. The import, export, or diversion of water within and between watersheds has been shown to affect a wide variety biotic and abiotic processes as a result of changes in the quantity and timing of surface water discharge and other aspects of the hydrologic regime (Taylor 1982; Kondolf et al. 1987; Stromberg and Patten 1990; Petts 1996; Davies, Thoms, and Meador 1992)

This indicator was measured as the percent of a riparian reach drainage basin from which surface water was imported, exported, or diverted on a continuous or periodic basis. In the case of imported water, the area of the watershed from which water was being imported was added to the area of the riparian reach drainage basin receiving water prior to calculating the percentage of the drainage basin that contributed to import. Using the ArcView GIS theme of riparian reaches and USGS 7.5 minute topographic map images the area below import, export, or diversion points were calculated and summed across the drainage basin to determine the metric value.

The reference condition was defined as <5% of the drainage basin of a riparian with surface water continuously or occasionally imported, exported, or diverted. Indicator scores were assigned based on the range of indicator values in Table 4.

Table 4. Range of indicator values for scaling the import, export, or diversion of water indicator

| Indicator Value Range | Score |
|---|---|
| <5% of drainage basin with import, export, or diversion of water | 5 |
| >5 and <15% of drainage basin with import, export, or diversion of water | 4 |
| >15 and <30% of drainage basin with import, export, or diversion of water | 3 |
| >30 and <50% of drainage basin with import, export, or diversion of water | 2 |
| >50% of drainage basin with import, export, or diversion of water | 1 |

### 5.5 Floodplain Interaction ($FI_{RR}$)

Floodplain Interaction indicates of the degree to which the overbank hydrologic connection between the bankfull channel and the active floodplain and terraces of the riparian ecosystem has been lost in a riparian reach. The lost connection could be a result of levees, channelization, or channel incision. Many of the characteristics and processes of riparian ecosystems are dependent on periodic hydrologic interaction between the stream channel and the floodplain. When the hydrologic connection is lost, the physical and biological characteristics of the riparian ecosystem change, regardless of the reason.

This indicator was measured as the percent of the main stem channel through a riparian reach that was physically disconnected from the floodplain as a result of culturally accelerated channel erosion/incision, channel improvements, or levees. An incised main stem channel in which the active floodplain has been reestablished within the incised channel through normal fluvial processes was not considered disconnected (Keller 1972). When one side of the channel was disconnected from the floodplain, 50% of the stream channel was considered disconnected from the floodplain and terraces. When both sides of the channel were disconnected from the floodplain, 100% of the stream channel was considered disconnected. Aerial photography and field observations were used to estimate the value of the metric.

The reference condition was defined as <5% of the main stem stream channel disconnected from the floodplain. Indicator scores were assigned based on the range of indicator values in Table 5.

Table 5. Range of indicator values for scaling the floodplain interaction indicator

| Indicator Value Range | Score |
|---|---|
| <5% of main stem stream channel disconnected from the floodplain | 5 |
| >5 and <15% of main stem stream channel disconnected from the floodplain | 4 |
| >15 and <30% of main stem stream channel disconnected from the floodplain | 3 |
| >30 and <50% of main stem stream channel disconnected from the floodplain | 2 |
| >50% of main stem stream channel disconnected from the floodplain | 1 |

### 5.6 Sediment Regime ($SR_{RR}$)

Sediment Regime indicates the degree to which the sediment dynamics in the main stem channel of a riparian reach are in equilibrium with respect to the supply of sediments from upstream sources and erosion and deposition processes within the channel. A variety of cultural activities can alter sediment dynamics and/or channel geometry. These types of changes include channel erosion due to physical disturbance, channel incision and head-cutting due to the alteration of slope, channel aggregation due structures that impede flow (i.e., weirs, drop structures, culverts), and irrigation diversions (Kondolf et al. 1987).

This indicator was assigned a score by matching field observations to the descriptions in Table 6. The reference condition was defined as exhibiting a sediment regime in equilibrium with respect to supply, erosion, and deposition processes, and not affected by cultural alteration.

### 5.7 Land Use / Land Cover in Drainage Basin ($LULC_{RRDB}$)

Land use / land cover (LULC) indicates the way in which a tract of land is utilized, has been developed, or the physiognomic class of vegetation. For example, a tract of land that is used to produce row crops is assigned an agricultural LULC, golf courses and parks are assigned to a recreational or open space LULC, urban areas are typically assigned to a residential, industrial, or commercial LULC. Lands supporting natural vegetation communities (i.e., chaparral versus pasture) are assigned to a shrub, forest, or grassland LULC. A variety of LULC classifications have been developed over the years. Today however, the reference to LULC usually implies The USGS classification of LULC (Anderson et al. 1976) or a similar, but more detailed regional variations of this classification. This type of LULC classification is typically developed through the interpretation of aerial photographs or the analysis of other remote sources of thematic information (USGS 1990).

Table 6. Description of Conditions for assigning sediment regime indicator score

| Description of Conditions | Score |
|---|---|
| *Movement of sediment in the channel is in equilibrium in terms of supply, erosion, and deposition processes that reflect the culturally unaltered condition.* On higher-order streams there are alternating point bars; bank erosion occurs, but is stabilized and moderated by vegetation; and channel width, form, and floodplain area is consistent through the reach. In low-order streams with bedrock control, some of these indicators may not be apparent, but overall bank and hillslope erosion is moderated by vegetation, and there are no apparent culturally induced catastrophic failures. | 5 |
| *Movement of sediment in the channel is in equilibrium with the current hydrologic regime, as opposed to a culturally unaltered condition, and exhibits an overall balance in terms of erosion and deposition processes.* On higher-order streams there are alternating point bars; bank erosion occurs, but is stabilized and moderated by vegetation; and channel width, form, and floodplain area is consistent through the reach. In low-order streams with bedrock control, some of these indicators may not be apparent, but overall bank and hillslope erosion is moderated by vegetation, and there are no apparent culturally induced catastrophic failures. | 4 |
| *Sediment disequilibrium minor and localized within the reach. This includes small, localized areas of bank protection, slumping, or encroachment on the floodplain and channel.* This condition class also includes previously disrupted reaches on a recovery trajectory, such as deeply entrenched streams where downcutting has been arrested by structural grade control, and there is sufficient room for lateral channel migration and establishment of a functional floodplain within the incised channel. | 3 |
| *Sediment erosion and deposition out of equilibrium.* Water inflow is sediment rich or poor, or accelerated bank erosion exists. Channel not actively incising, but extensive disequilibrium is evident. Typical indicators include extensive bank slumping (erosion events that exceed any moderating influence of native vegetation), active gullies feeding into the reach from adjacent hillslopes, shoaling of sediments rather than deposition in sorted lateral and mid-channel bars. Apparently stable channels should be placed in this category if there is evidence of regular mechanical disruption, such as bulldozing of the channel bottom and clearing of riparian vegetation to improve flood conveyance. | 2 |
| *Sediment dynamics within most of the reach are seriously disrupted.* This includes reaches where there is no significant storage or recruitment of sediment (i.e., reaches in underground tunnels/culverts, and reaches hardened with rock or concrete). It also includes reaches that are either actively incising or functioning as sediment traps (e.g., sediment basins). This also includes reaches that have been subject to recent changes likely to induce severe disequilibrium, such as extensive floodplain filling, change in slope, channel straightening, or other changes that are likely to cause channel downcutting during future high-flow events. | 1 |

Over the centuries, humans have modified the LULC of the natural landscape through intensive land management practices such as agriculture, forestry, and grazing, as well as

through industrialization and urbanization. The net effect of these activities has been a dramatic shift in the type and extent of LULC that occur around the world today, particularly in developed countries (Meyer and Turner 1992; Hannah et al. 1994.

A number of studies have related LULC to water quality. While they have consistently shown that the water quality decreases as natural LULC are culturally altered, they specific relationships and causative factors vary widely. For example, Hunsaker and Levine found that LULC changes in the watershed had the greatest effect on water quality, while Graf 1998 found that changes in LULC in the surrounding landscape had the greatest effect. The relationship between LULC and quantity and quality of surface water has been documented for a variety of wetland and aquatic systems (Brugham 1978; Ehrenfield 1983; Kuenzler 1986; Howarth et al. 1991; Ryan 1991; Williamson et al. 1992; Richards and Host 1994; Cooper 1995; Blair 1996; Wilber et al. 1996; Caruso and Ward 1998). In the western United States specifically, livestock grazing, agriculture, and urbanization have often been identified as contributors to increased surface water runoff and non-point sources of sediment, nutrients, and other classes of pollutants (Armour et al. 1991; Sedgwick and Knopf 1991; Charbonneau and Kondolf 1993; Bush and Smith 1995; Rothrock et al. 1998).

Four sub-indicators were used to measure the LULC indicator. Each of the sub-indices were measured as the percent of the drainage basin of a riparian reach with LULC types with the potential to increase the nutrient, pesticide, hydrocarbon, or sediment loading in downstream surface waters. Land use / land cover categories with the potential to increase these categories of pollutants are shown in Table 7. Using the ArcView GIS themes of riparian reach and LULC themes, the area of a drainage basin occupied by each LULC was determined for each sub-indicator. The area of LULC types with the potential to increase pollutants, hydrocarbons, nutrients, and sediment were then summed across the drainage basin and divided by the total drainage basin area to determine the sub-indicator value. The four sub-indicator values were averaged to determine the LULC indicator value.

The reference standard condition was defined as <5% of the watershed and surrounding landscape area with LULC types with the potential to increase nutrient, pesticide, hydrocarbon,

Table 7. Land Use / Land Cover (LULC) types

| LULC Code | LULC Description | Increase in LULC Type Increases Runoff | Increase in LULC Type Increases Nutrient Load | Increase in LULC Type Increases Pesticide Load | Increase in LULC Type Increases Hydrocarbon Load | Increase in LULC Type Increases Sediment Load | Increase in LULC Type Inhibits Animal Use / Movement |
|---|---|---|---|---|---|---|---|
| 11 | Residential | Yes | Yes | Yes | Yes | No | Yes |
| 12 | Commercial and Services | Yes | Yes | Yes | Yes | No | Yes |
| 13 | Industrial | Yes | Yes | Yes | Yes | No | Yes |
| 14 | Transportation/Commercial/Utilities | Yes | Yes | Yes | Yes | No | Yes |
| 15 | Industrial and Commercial | Yes | Yes | Yes | Yes | No | Yes |
| 16 | Mixed Urban and Built-Up | Yes | Yes | Yes | Yes | No | Yes |
| 17 | Other Urban or Built-Up | Yes | Yes | Yes | Yes | No | Yes |
| 21 | Cropland and Pastureland | Yes | Yes | Yes | No | Yes | Yes |
| 22 | Orchards/Vinyards/Nurseries | Yes | Yes | Yes | No | Yes | Yes |
| 23 | Confined Feeding Operations | Yes | Yes | No | No | Yes | Yes |
| 24 | Other Agricultural Land | Yes | Yes | Yes | No | Yes | Yes |
| 31 | Herbaceous Rangeland | Yes | Yes | Yes | No | Yes | Yes |
| 32 | Shrub and Brush Rangeland | Yes | Yes | Yes | No | Yes | No |
| 33 | Mixed Rangeland | Yes | Yes | Yes | No | Yes | Yes |
| 41 | Deciduous Forest Land | No | No | No | No | No | No |
| 42 | Evergreen Forest Land | No | No | No | No | No | No |
| 43 | Mixed Forest Land | No | No | No | No | No | No |
| 52 | Lakes | No | No | No | No | No | No |
| 53 | Reservoirs | No | No | No | No | No | No |
| 54 | Bays and Estuaries | No | No | No | No | No | No |
| 61 | Forested Wetlands | No | No | No | No | No | No |
| 62 | Nonforested Wetlands | No | No | No | No | No | No |
| 71 | Dry Salt Flats | No | No | No | No | No | No |
| 72 | Beaches | No | No | No | No | No | No |
| 73 | Sandy Areas (non-Beach) | No | No | No | No | No | No |
| 74 | Exposed Rock | No | No | No | No | No | No |
| 75 | Strip Mines | Yes | Yes | Yes | Yes | Yes | Yes |
| 76 | Transitional Areas | Yes | Yes | Yes | Yes | Yes | Yes |

or sediment loading in surface waters downstream. Indicator scores were assigned based on the range of indicator values in Table 8.

Table 8. Range of indicator values for scaling the land use / land cover indicator

| Indicator Value Range | Score |
|---|---|
| <5% of watershed / landscape with LULC types that increase N/P/H/S | 5 |
| >5 and <15% of watershed / landscape with LULC types that increase N/P/H/S | 4 |
| >15 and <30% of watershed / landscape with LULC types that increase N/P/H/S | 3 |
| >30 and <50% of watershed / landscape with LULC types that N/P/H/S | 2 |
| >50% of watershed / landscape with LULC types that increase N/P/H/S | 1 |

### 5.8 Area of Native Riparian Vegetation ($NRV_{RR}$)

Area of Native Riparian Vegetation indicates the degree to which native riparian vegetation communities occupy the floodprone area of the main stem channel through a riparian reach. Much has been written about the importance of native riparian vegetation communities in the support of specific faunal groups such as amphibians (Brode and Bury 1984), birds (Hendricks and Rieger 1989), and fauna in general (Hubbard 1977; Faber et al. 1989; Knopf et al. 1988). This indicator was measured as the percent of floodprone area along the main stem channel of the riparian reach occupied by native riparian vegetation communities. Under culturally unaltered conditions, a complex interaction of many factors such as the size of the watershed, discharge, channel geometry, substrate type, and slope determine the size of the area that typically supports riparian vegetation. In general however, as stream orders increase, the width of the bankfull channel increases, and the size of the area supporting riparian vegetation increases. Floodprone area represents a scaled metric that can be applied consistently in different stream orders throughout a watershed. Floodprone area was determined in the field by projecting the elevation corresponding to two times the maximum depth of the bankfull channel until it intersected the surface of theadjacent floodplain / terrace on both sides of the main stem channel (Rosgen 1996; 5-20). The percent of floodprone area occupied by native riparian vegetation was estimated based on field observations, aerial photographs, and riparian vegetation communities mapped by Lichvar (2000).

The reference condition was defined as >95% of the floodprone width of the main stem channel through the riparian reach occupied by native riparian vegetation communities. Indicator scores were assigned based on the range of indicator values in Table 9.

Table 9. Range of indicator values for scaling the native riparian vegetation indicator

| Indicator Value Range | Score |
|---|---|
| >95% of floodplain occupied by riparian vegetation communities | 5 |
| <95 and >85 and % of floodplain occupied by riparian vegetation communities | 4 |
| <85 and >70 and % of floodplain occupied by riparian vegetation communities | 3 |
| <70 and >50 and % of floodplain occupied by riparian vegetation communities | 2 |
| <50% of floodplain occupied by riparian vegetation communities | 1 |

### 5.9 Riparian Corridor Continuity ($RCC_{RR}$ / $RCC_{DB}$)

Riparian Corridor Continuity indicates the degree to which the main stem channel of a riparian reach exhibits an uninterrupted vegetated riparian corridor. Riparian ecosystems typically form a relatively continuous corridor along the stream channel and floodplain. Intact vegetated corridors allow animals to move to locations throughout a watershed on a daily, seasonal, or annual basis (La Polla and Barrett 1993; Machtans et al. 1993; Naiman et al. 1993 and 1996), but see Simberloff et al. (1992). Gaps in the continuous riparian corridor can occur as a result of natural fluvial processes during large magnitude events (Hawkins et al. 1997). However, gaps are more frequently created as a result of cultural alterations such as roads, power and pipeline corridors, agriculture activities, and urban/industrial development.

This indicator was measured at the riparian reach scale as the percent of floodprone area along the main stem channel of the riparian reach occupied by native and non-native vegetation communities with adequate height and structure to allow faunal movement. For example, annual grassland with no shrub or tree component was considered to represent a corridor gap. The difference between this indicator and Area of Native Riparian Vegetation was that for the RCC - RR indicator, the vegetation corridor could be composed of native or non-native riparian species, whereas for the ANRV indicator, only native riparian vegetation communities were considered. The percent of floodprone area occupied by native riparian vegetation was estimated based on field observations, aerial photographs, and riparian vegetation communities mapped by Lichvar (2000). At the drainage basing scale, Riparian Corridor Continuity was calculated as the

weighted average of the percent of Riparian Corridor Continuity for all riparian reaches in the drainage basin of the riparian reach. In other words,

$$\sum_{i=1}^{i_n}\left[\frac{RCC_{RR} * \left(ML_{RR} / ML_{DB}\right)}{100}\right] \qquad (2)$$

Where: $RRC_{RR}$ = % of main stem in a riparian reach with vegetation corridor gaps

$ML_{RR}$ = Length of main stem channel in a riparian reach

$ML_{DB}$ = Length of main stem channel of all riparian reaches in drainage basin

The reference condition was defined as <5% of the floodplain of the main stem channel of the riparian reach occupied with riparian vegetation communities. Indicator scores were assigned based on the range of indicator values in Table 10.

Table 10. Range of indicator values for scaling the riparian corridor continuity indicators

| Indicator Value Range | Score |
|---|---|
| <5% of riparian reach with gaps / breaks due to cultural alteration | 5 |
| >5 and <15% of riparian reach with gaps / breaks due to cultural alteration | 4 |
| >15 and <30% of riparian reach with gaps / breaks due to cultural alteration | 3 |
| >30 and <50% of riparian reach with gaps / breaks due to cultural alteration | 2 |
| >50% of riparian reach with gaps / breaks due to cultural alteration | 1 |

### 5.10 Land Use / Land Cover at Riparian Ecosystem Boundary ($BND_{RR}$)

Land Use / Land Cover at Riparian Ecosystem Boundary indicates the presence of cultural features at the boundary of the riparian ecosystem that are likely to inhibit the normal movement of fauna between riparian and adjacent upland habitats. Land use / land cover at the boundary of the riparian ecosystem plays an important role in determining the ability of animals to move freely between riparian and adjacent upland ecosystems on a daily or seasonal basis (Petersen et al. 1992; Vought et al. 1994, Statzner et al. 1997; Vought et al. 1994; Osborne and Kovacic 1993). Under natural conditions, riparian vegetation transitions gradually to native upland vegetation at the edge of the riparian ecosystem. A variety of cultural activities replace these native or naturalized vegetation communities with agriculture, urban/industrial, transportation

corridors or other types of LULC that reduce the likelihood the animals can move freely between the riparian ecosystem and adjacent uplands.

This indicator was measured using the ArcView GIS themes of riparian reach and LULC. A one-meter wide buffer strip was created at the riparian ecosystem boundary to simulate a boundary condition. The percent of this area occupied by each LULC type was determined, and the LULC types considered to inhibit faunal movement in Table 8 were used to determine the value of the metric.

The reference condition was defined as <5% of the riparian ecosystem boundary composed of LULC types that inhibit faunal movement. Indicator scores were assigned based on the range of indicator values in Table 11.

Table 11. Range of indicator values for scaling the riparian ecosystem boundary indicator

| Indicator Value Range | Score |
|---|---|
| <5% of the riparian boundary inhibits faunal movement | 5 |
| >5 and <15% of the riparian boundary inhibits faunal movement | 4 |
| >15 and <30% of the riparian boundary inhibits faunal movement | 3 |
| >30 and <50% of the riparian boundary inhibits faunal movement | 2 |
| >50% of the riparian boundary inhibits faunal movement | 1 |

### 5.11 Land Use / Land Cover in Upland Buffer ($BUF_{RR}$)

Land Use / Land Cover in Upland Buffer indicates the degree to which the LULC in a buffer zone has been culturally altered. Land Use / Land Cover -Upland Buffer differs from the Land Use / Land Cover - Riparian Reach Boundary indicator in that it is concerned with LULC in the entire adjacent upland landscape and not just at the boundary between the riparian ecosystem and the adjacent upland. Land use / land cover in upland areas adjacent to riparian ecosystems are important because of their ability to support the life requirements of a variety of native species. Under reference conditions the upland buffer consists of native vegetation communities. A variety of cultural activities replace these native or naturalized vegetation communities with agriculture, urban/industrial, transportation corridors or other types of land use. Changes in LULC in the buffer also have the potential to affect the rate at which water and sediment moves toward riparian areas from the uplands (Peterjohn and Correll 1984, 1986; Osborne and Kovacic 1993; Barling and Moore 1994).

This indicator was measured using field observation, aerial photographs, and the ArcView GIS themes of riparian reach and LULC themes. A buffer of 100 m width, or until an adjacent 2nd order or higher watershed boundary was encountered, was established around the riparian reach. The percent of the buffer area occupied by each LULC type was determined, and the LULC types considered to inhibit faunal use in Table 8 were used to determine the value of the indicator.

The reference condition was defined as <5% of the upland buffer with LULC types representing cultural alteration. Indicator scores were assigned based on the range of indicator values in Table 12.

Table 12. Range of indicator values for scaling the riparian ecosystem upland buffer indicator

| Metric Value Category | Score |
|---|---|
| <5% of the buffer zone with culturally altered LULC | 5 |
| >5 and <15% of the buffer zone with culturally altered LULC | 4 |
| >15 and <30% of the buffer zone with culturally altered LULC | 3 |
| >30 and <50% of the buffer zone with culturally altered LULC | 2 |
| >50% of the buffer zone with culturally altered LULC | 1 |

## 6.0 Assessment Procedure (Methods)

The assessment of riparian ecosystem integrity in the San Juan/San Mateo Watersheds was conducted by completing the following tasks described in the sections below:

Task 1: Identification of riparian reach assessment units

Task 2: Characterization of riparian reaches

Task 3: Assessment of indicators

Task 4: Assigning indicator scores and calculation of indices

Task 5: Archiving of information

### 6.1 Identification of Riparian Reach Assessment Units

On non-headwater riparian reaches (i.e., riparian reaches with other riparian reaches upstream) the longitudinal (i.e., upstream / downstream) boundaries of a riparian reach corresponded to changes in stream gradient or channel morphology resulting from geological control (e.g. knick points), tributaries / distributaries, artificial grade control structures, or other

features related to cultural alteration. On headwater reaches, the upstream end of main stem channel of headwater riparian reach always included third order streams (Strahler 1952, 1957) as mapped by Lichvar (2000), and in many cases the upstream end included second order streams. The factors that determined the upstream extent of the riparian reach were stream density, accessibility, and the extent of a riparian vegetation component. Lateral boundaries of riparian reaches corresponded to either an estimate of the 100-year flood elevation contour, the extent of identifiable historic alluvial terraces, or the base of valley wall or artificial structure. In all cases the riparian reach included Riparian Zones 1 and 2 discussed above. Each riparian reach was assigned pneumonic identifier for display and digital manipulation purposes. The local drainage area of each riparian reach was digitized in ArcView using a digital raster graphic (DRG) background. The local drainage of a riparian reach was combined with the local drainages of all upstream riparian reaches to define the drainage basin of the riparian reach (Figure 2).

Preliminary riparian reach, local drainage area, and drainage basin boundaries were mapped on the basis of initial field reconnaissance, aerial photos, and maps. These preliminary boundaries were modified throughout the study based on field visits to each riparian reach, and the WoUS maps developed by Lichvar (2000). Polygons representing the riparian reach, local drainage area, and drainage basin boundaries were constructed in ArcView (see Appendix A).

### 6.2 Characterization of Riparian Reaches

Characterization of riparian reaches was accomplished during a site visit to each reach. Several reaches in the roadless areas such as Upper Trabuco Canyon were not visited due to difficult accessibility and time constraints. For these reaches, the characterization was completed to the extent possible using aerial photographs and topographic maps. Table 13 provides a listing of the information collected as part of riparian reach characterization. A complete listing of the information collected for all riparian reaches as part of the characterization effort is provided in Table 15 in Appendix A.

The general strategy during a site visit was to begin at the downstream end of the riparian reach and conduct a walking reconnaissance of the main stem channel through the riparian reach. On longer reaches we drove to representative sections of the riparian reach and conduct separate walking reconnaissance. On headwater reaches the walking reconnaissance included at least the

Table 13. List of data for characterizing riparian reaches and indicators

| Field Description | Method |
|---|---|
| Riparian Reach ID | Field |
| Drainage Basin | Field |
| USGS 7.5 Minute Topograpic Quad | GIS |
| Main stem Downstream End Coordinates (UTM) | GIS |
| Main stem Upstream End Coordinates (UTM) | GIS |
| Size of Mapped Riparian Ecosystem in Riparian Reach Local Drainage (ha) | GIS |
| Size of Mapped Riparian Ecosystem in Riparian Reach Drainage Basin (ha) | GIS |
| Size of Riparian Reach Local Drainage (RRLD) (ha) | GIS |
| Length of RRLD Perimeter (m) | GIS |
| Size of Riparian Reach Drainage Basin (RRDB) Area (ha) | GIS |
| Valley Type (Rosgen) | Field |
| Valley Length (m) | Field / GIS |
| Valley Width (m) | Field / GIS |
| Main stem Downstream End Elevation (m) | GIS |
| Main stem Upstream End Elevation (m) | GIS |
| Valley Slope (%) (Estimated From 7.5 Minute Topo) | Calculated |
| Engineered Channel Type or Rosgen Stream Type | Field |
| Main stem Channel Length in RRLD (m) (Smith) | GIS |
| Main stem Channel Length in RRDB (m) (Smith) | GIS |
| Main stem and Tributary Channel Length in RRLD (m) (Lichvar) | GIS |
| Main stem and Tributary Channel Length in RRDB (m) (Lichvar) | GIS |
| Main stem Channel Length / Main stem Channel and Tributary Channels Length | Calculated |
| Drainage Density | Calculated |
| Channel Slope | Calculated |
| Sinuosity | Calculated |
| Bankfull Width (ft) | Field |
| Bankfull Width (m) | Calculated |
| Floodprone Width (ft) | Field |
| Floodprone Width (m) | Calculated |
| Bankfull Maximum Depth (in) | Field |
| Bankfull Maximum Depth (cm) | Calculated |
| Bankfull Mean Depth (in) | Field |
| Bankfull Mean Depth (cm) | Calculated |
| Bankfull Cross-Sectional Area ($m^2$) | Calculated |
| Width / Depth Ratio | Calculated |
| Entrenchment Ratio | Calculated |
| Natural Channel Substrate Bedrock / Boulder (%) | Field |
| Natural Channel Substrate Cobble (%) | Field |
| Natural Channel Substrate Gravel (%) | Field |
| Natural Channel Substrate Sand (%) | Field |

Table 13 cont. List of data for characterizing riparian reaches and indicators

| Field Description | Method |
|---|---|
| Natural Channel Substrate Silt / Clay (%) | Field |
| Indicator 1 % of Channel in RRLD with Altered Hydraulic Conveyance | Field |
| Indicator 2 % of Channel in RRDB with Altered Hydraulic Conveyance | GIS |
| Indicator 3a % of Flood Prone Area with Native Riparian Vegetation | Field |
| Indicator 3b % of Flood Prone Area with Native Riparian Vegetation | GIS |
| Indicator 4 % of Floodplain Present and not Isolated from Channel | Field |
| Indicator 5 % of Channel with Perennialized Stream Flow | Field |
| Indicator 6 % of Flood Prone Area in the RRLD with Riparian Corridor Breaks | Field |
| Indicator 7 % of Flood Prone Area in the RRDB with Riparian Corridor Breaks | Field |
| Indicator 8a % of Buffer (100m) with Culturally Altered LULC types | Field |
| Indicator 8b % of Buffer (100m) with Culturally Altered LULC types | GIS |
| Indicator 9 Sediment Regime Condition Index of Channel in RRLD | Field |
| Indicator 10 % of RRDB with Surface Water Imported, Exported or Diverted | GIS |
| Indicator 11 % of LULC Contributing to Nutrient Increase in Surface Waters | GIS |
| Indicator 12 % of LULC Contributing to Pesticide Increase in Surface Waters | GIS |
| Indicator 13 % of LULC Contributing to Hydrocarbon Increase in Surface Waters | GIS |
| Indicator 14 % of LULC Contributing to a Sediment Increase in Surface Waters | GIS |
| Indicator 15 % of RRLD with a Culturally Altered Boundary | GIS |
| Indicator 16 % of RRDB with Surface Water Retention | GIS |

lower-third of the main stem channel of the riparian reach. Time constraints precluded conducting a walking reconnaissance of the entire main stem channel of headwater reaches in several roadless areas. In these situations, field observations were supplemented with interpretations based on aerial photographs.

After reconnaissance through a riparian reach, a decision was made to either retain the preliminary riparian reach boundaries, or to further divide the riparian reach into two or more riparian reaches. Based on the observations made during reconnaissance, a representative portion of the riparian reach was selected and a riparian reach characterization data sheet was completed. The characterization included notes on the species and location of the dominant vegetation in terms of geomporphic features (i.e., bank full channel, floodplain, terrace), measurement of channel characteristics, general field notes about the nature of the riparian reach, and indicator values for those integrity indicators measured in the field.

### 6.3 Assessment of Indicators

Indicators were assessed using a combination of fieldwork and spatial analysis in ArcView. Indicators assessed based on field observations included:

- Altered Hydraulic Conveyance ($AHC_{RR}$ and $AHC_{RRDB}$ )
- Perennialized Stream Flow ($PSF_{RR}$)
- Area of Native Riparian Vegetation ($NRV_{RR}$)
- Riparian Corridor Continuity ($RCC_{RR}$)
- Floodplain Interaction ($FI_{RR}$)
- Sediment Regime ($SR_{RR}$)

Indicators and sub-indicators measured using ArcView GIS included:

- Altered Hydraulic Conveyance – Drainage Basin ($AHC_{RRDB}$)
- Perennialized Stream Flow ($PSF_{RRDB}$)
- Riparian Corridor Continuity ($RCC_{RRDB}$)
- Surface Water Retention ($SWR_{RRDB}$)
- Import, Export, or Diversion of Surface Water ($IED_{RRDB}$)
- Land Use / Land Cover ($LULC_{RRDB}$)
    - Land Use / Land Cover – Nutrient Increase ($LULC\text{-}N_{RRDB}$)
    - Land Use / Land Cover – Pesticide Increase ($LULC\text{-}P_{RRDB}$)
    - Land Use / Land Cover – Hydrocarbon Increase ($LULC\text{-}H_{RRDB}$)
    - Land Use / Land Cover – Sediment Increase ($LULC\text{-}S_{RRDB}$)
- Land Use / Land Cover at Riparian Reach Boundary ($BND_{RR}$)
- Land Use / Land Cover in Upland Buffer ($BUF_{RRLD}$)

Information on the specific procedure used to measure each indicator is given in the Assessment Indicators section above.

Based on field observation and/or spatial analysis, each indicator was assigned a value representing the percent deviation of the indicator from the reference condition in that reach. For example, if the main stem of a riparian reach was completely channelized, an indicator value of 100 was assigned to the Altered Hydraulic Conveyance – Riparian Reach indicator. The assignment of values to indicators was based on an assumed, relative, categorical relationship

defined between indicators and assessment endpoints. The assumption was that an increase in the deviation from the reference condition represented an equivalent decrease in the level of riparian ecosystem integrity in terms of the specific indicator. For example, in comparing two riparian reaches in terms of the land use / land cover indicator, the riparian reach with the larger percentage of urban land use / land cover in the drainage basin would be assumed to have lower integrity, at least in terms of the contribution of the indicator to endpoint integrity. This approach to scaling indicators to reference condition, and by implication to endpoint integrity, was possible because of the way indicators were defined (i.e., always measurable as a percent deviation from reference condition), and based on information from published studies, field data and observations, and professional judgment as outlined the Assessment Endpoints and Assessment Indicators sections above.

### 6.4 Assigning Indicator Scores and Calculation of Indices

To calculate integrity indices, indicator values (i.e., percent deviation from reference condition 0-100) were converted to scores. Indicator values were assigned a score of 1-5 based on the relationship between indicator values and scores defined in Tables 1-12. Using the altered hydrologic conveyance indicator as an illustration (see Table 1), if the value of the altered hydrologic conveyance indicator for a riparian reach was <5%, a score of 5 was assigned to that riparian reach. If the value of the altered hydrologic conveyance indicator for a riparian reach was >30% and <50%, a score of 2 was assigned to that riparian reach. A score of 5 represented close concurrence with the reference condition, and consequently a high level of integrity. A score of 1 represented a deviation of 50% or more the reference condition, and consequently a low level of integrity.

Initial category ranges for indicator values were based on the natural groupings of the data collected during the project, and the subjective integration of numerous field observations relating indicator values to endpoint integrity. Testing of other category ranges (i.e., correlation analysis using quartiles and quintiles) showed no significant change in the relationship between riparian reaches in terms of either indicator scores or endpoint indices. Thus, initial category ranges were retained.

Hydrologic, water quality, and habitat integrity indices were calculated in the spreadsheet by summing the scores of the indicators associated with hydrologic, water quality, and habitat integrity as discussed above. Individual indicator scores and summary indices were presented in tabular form in the spreadsheet and spatially in ArcView.

### 6.5 Archiving of Information

All of the information and data collected during the characterization and assessment of riparian reaches (discussed above), as well as results derived from this information (discussed below), were archived in an Excel spreadsheet, ArcView project file, and an Access database format. Appendices A, B, and C discuss the archiving of this information, data, and results in each of these formats respective

## 7.0 Results and Discussion

Three hundred and sixty riparian reaches were identified in the San Juan and San Mateo Creek Watersheds. The size of the riparian reach local drainages ranged from 4 to 1799 hectares with a mean of 142 hectares, and the size of riparian reach drainage basins ranged from 6 to 45798 hectares with a mean of 1941 hectares. The length of the main stem channels through the riparian reach ranged from 0 to 8951 meters with a mean of 1500 meters. The wide ranges in these characteristics primarily reflect the extreme difference in the size of riparian reaches identified in hetreogenous urban versus more homogenous natural landscapes. The minimum, maximum, and mean of indicator values, and the frequency of indicator scores for all riparian reaches is summarized in Table 14.

The range of values for the endpoint indices (i.e., sum of relevant indicator scores) for hydrologic integrity was 6 - 26 with a mean of 22 out of a possible 30, for water quality integrity was 13 - 41 with a mean of 32 out of a possible 45, and for habitat intergrity was 5 - 25 with a mean of 20 out of a possible 30**.** Figures 3, 4, and 5 show the distribution of endpoint indices

Table 14. Descriptive statistics for indicator values and indicator scores

| Indicator | Min Value | Max Value | Mean Value | Frequency | | | | |
|---|---|---|---|---|---|---|---|---|
| | | | | Score 5 | Score 4 | Score 3 | Score 2 | Score 1 |
| $AHC_{RR}$ | 0 | 100 | 20 | 283 | 0 | 2 | 1 | 7 |
| $AHC_{RRDB}$ | 0 | 100 | 18 | 254 | 22 | 9 | 12 | 63 |
| $ARV_{RR}$ | 0 | 100 | 80 | 250 | 4 | 27 | 19 | 60 |
| $FI_{RR}$ | 0 | 100 | 67 | 234 | 1 | 4 | 6 | 117 |
| $PSF_{RR}$ | 0 | 100 | 20 | 283 | 1 | 4 | 1 | 71 |
| $RCC_{RR}$ | 0 | 100 | 24 | 241 | 9 | 19 | 17 | 74 |
| $RCC_{RRDB}$ | 0 | 100 | 29 | 239 | 28 | 13 | 16 | 64 |
| $BUF_{RRLD}$ | 0 | 100 | 49 | 133 | 14 | 16 | 36 | 161 |
| $SR_{RR}$ | NA | NA | NA | 111 | 64 | 86 | 37 | 62 |
| $IED_{RRDB}$ | 0 | 0 | 0 | 360 | 0 | 0 | 0 | 0 |
| $LULC_{N}$ | 0 | 100 | 43 | 79 | 27 | 42 | 53 | 159 |
| $LULC_{P}$ | 0 | 100 | 36 | 112 | 24 | 45 | 55 | 124 |
| $LULC_{H}$ | 0 | 100 | 15 | 232 | 37 | 22 | 20 | 49 |
| $LULC_{S}$ | 0 | 100 | 30 | 98 | 39 | 76 | 54 | 93 |
| $BND_{RR}$ | 0 | 100 | 28 | 178 | 29 | 28 | 27 | 98 |
| $SWR_{RRDB}$ | 0 | 100 | 10 | 283 | 19 | 12 | 14 | 32 |

across all riparian reaches for hydrologic, water quality, and habitat integrity respectively. In general the index values exhibited a relatively wide and even spread across the possible range of index values. These results can be interpreted as evidence that the indicators were scaled appropriately, and sensitive enough to distinguish varying degrees of hydrologic, water quality,

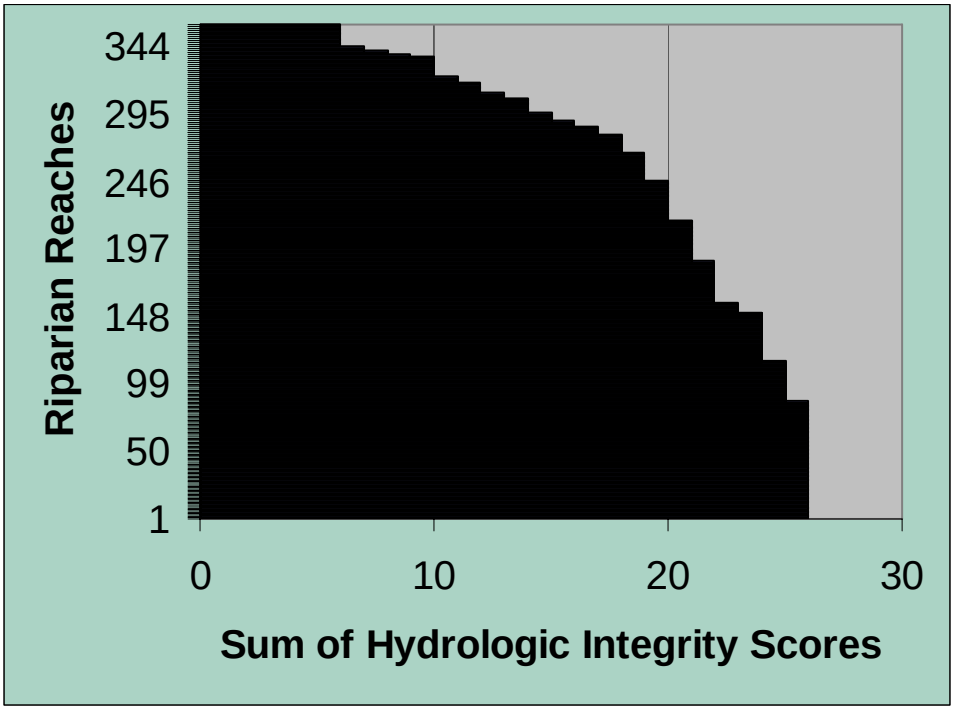


Figure 3. Index values for hydrologic integrity for all riparian reaches

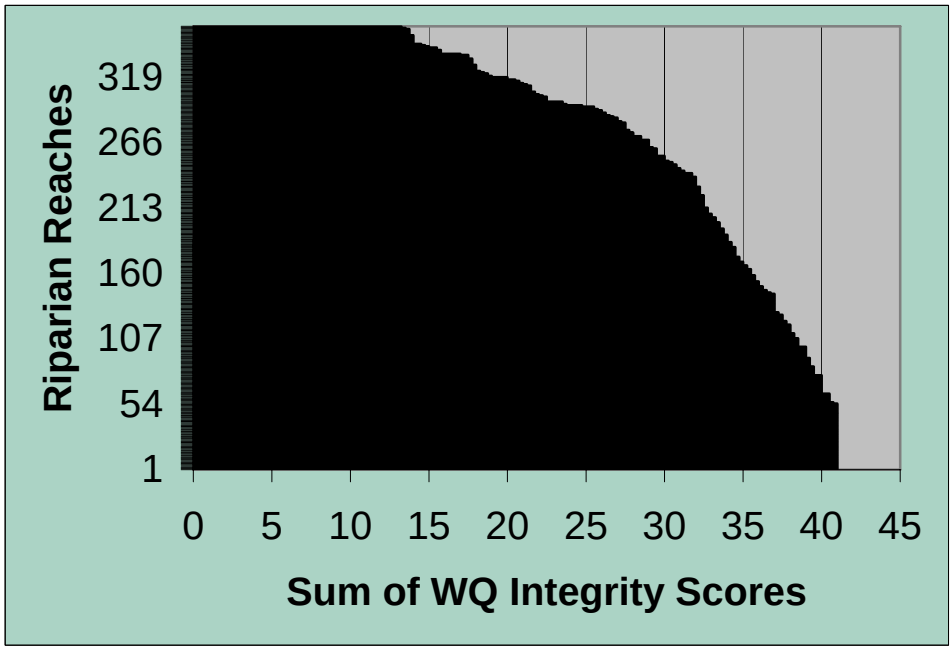


Figure 4. Index values for water quality integrity for all riparian reaches

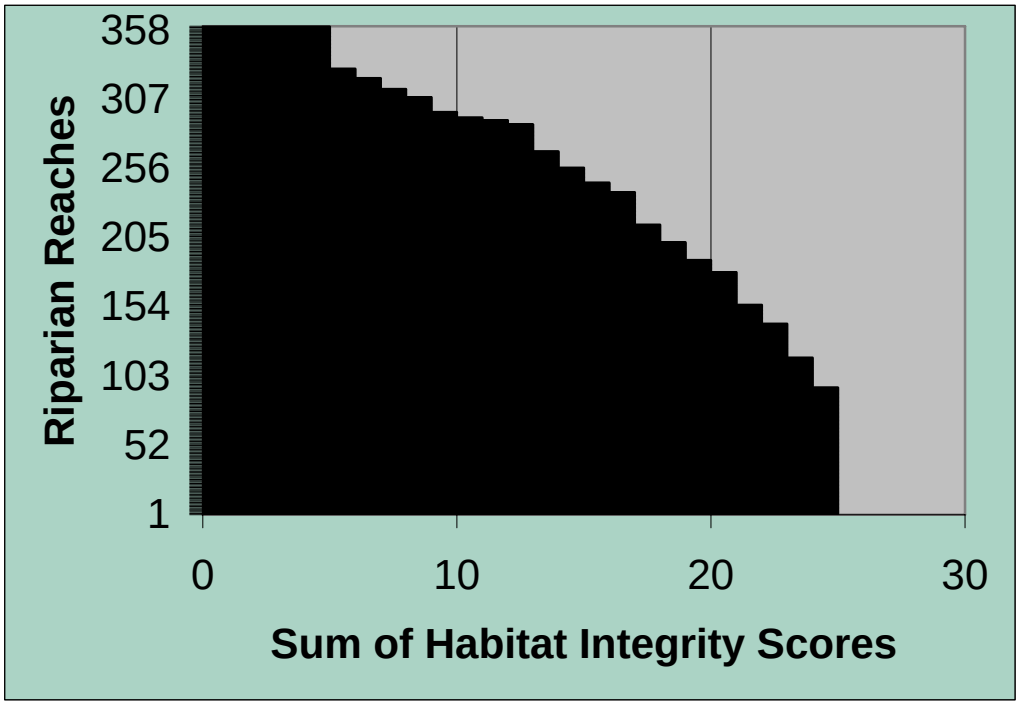


Figure 5. Index values for habitat integrity for all riparian reaches

and habitat integrity. This is encouraging given that the results fit our perception of riparian ecosystem integrity in the watershed based on extensive field-work and observations. Ultimately, however, the only way to increase confidence in the integrity indices is through testing an verification with more quantitative models of hydrologic, water quality, and habitat integrity.

Because of the extensive amount of data collected for each riparian reach and the inherently spatial nature of the data, the Excel spreadsheet and ArcView project file are much better formats for reviewing the results in terms of the three tasks required to meet the project objective (see Project Objective section above). Consequently, the objective here will be to provide examples and illustrations to acquaint readers with the way in which results can be presented and summarized in the database, spreadsheet, and ArcView formats.

Figure 6 illustrates the summary report available for each riparian reach in the database (see Appendix C). This summary report is an ideal starting point for reviewing results because it provides a good overview of a riparian reach in terms of its characteristics and indicator scores.

# San Juan/San Mateo Watersheds
## Riparian Reach Characterization and Functional Assessment Summary

### General Information
Drainage Basin: Bell Canyon
Riparian Reach ID: BE-07
USGS 7.5 Minute Topo: Canada Gobanadora
UTM Coordinates Downstream End:
11S 447150mE 3716058mN
UTM Coordinates Upstream End:
11S 447260mE 3716882mN
Size of Riparian Reach: 64 ha
Size of Drainage Basin: 117 ha
Area of Riparian Ecosystem: 3.2 ha

### Channel Characteristics
Channel Type or Rosgen Stream Type if Natural Channel: C3
Length of Main stem Channel Through Reach: 1090 m
Channel Substrates (Natural Channels Only):
- % Bedrock or Boulder: 0
- % Cobble: 10
- % Gravel: 20
- % Sand: 60
- % Silt / Clay: 10

Channel Geometry in Representative Section of Lower Portion of Reach:
- Bankfull Width: 11.3 m
- Flood Prone Width: 19.5 m
- Mean Bankfull Depth: 38.1
- Bankfull Cross-Sectional Area: 1.7 $m^2$

### Indicators of Functional Integrity
% of Reach with Altered Hydraulic Conveyance: 0
% of Drainage Basin with Altered Hydraulic Conveyance: 16
% of Floodplain Removed or Isolated from Channel: 0
% of Channel with Perennial Flow Due to Supplementary Sources: 0
Sediment Regime Condition Index: 2
% of Drainage Basin Surface Water Imported, Exported, or Diverted: 0
% of Drainage Basin affected by Surface Water Storage Structures: 93
% of Drainage Basin with Land Uses that increase surface water nutrients: 93
% of Drainage Basin with Land Uses that increase surface water pesticides: 93
% of Drainage Basin with Land Uses that increase surface water hydrocarbons: 93
% of Drainage Basin with Land Uses that increase surface water sediments: 93
% of Flood Prone Area in Reach Functioning as Corridor Breaks: 0
% of Flood Prone Area in Drainage Basin Functioning as Corridor Breaks: 0

Figure 6. Page 1 of a database summary report

**Indicators of Functional Integrity cont.**
% of Riparian Ecosystem Boundary with Culturally Altered Land Uses: 100
% of Riparian Ecosystem Buffer (100 m) with Culturally Altered Land Uses: 100
% of Flood Prone Area supporting Native Riparian Vegetation: 100

**Indicator Scores**

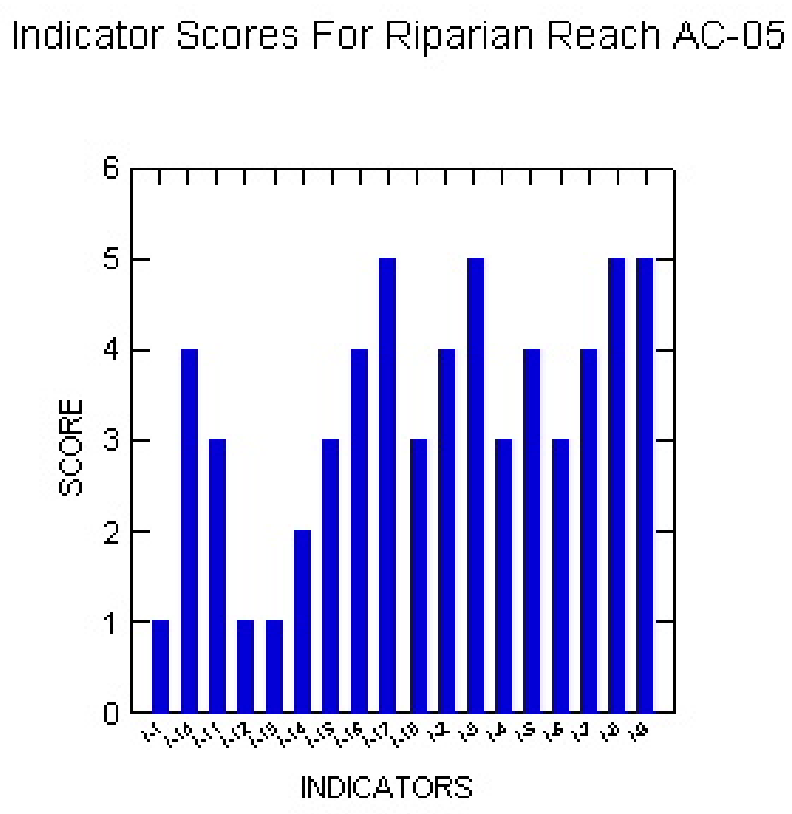


Figure 6 cont. Page 2 of a database summary report

Figure 7 is an ArcView layout showing the location of riparian reaches in the San Juan/San Mateo Watersheds. Each polygon in the figure represents the local drainage of a riparian reach. Labels are the codes assigned to each riparian reach and used to identify riparian reaches in the spreadsheet, ArcView project, and database.

Figure 7. Location of riparian reaches as represented by labeled local drainages

Figure 8 shows a portion of the San Juan/San Mateo Watersheds in the vicinity of Borrego and Serrano Creeks. In the figure, brown polygons again represent the local drainage of a riparian reach. The riparian ecosystem in each reach consists of blue lines and colored polygons representing wetland ratings from Lichvar (2000). This type of display can be quickly presented for any portion of the San Juan/San Mateo Watersheds in ArcView. Attachment 1 is a fold-up map that shows this same view for the entire San Juan/San Mateo Watersheds.

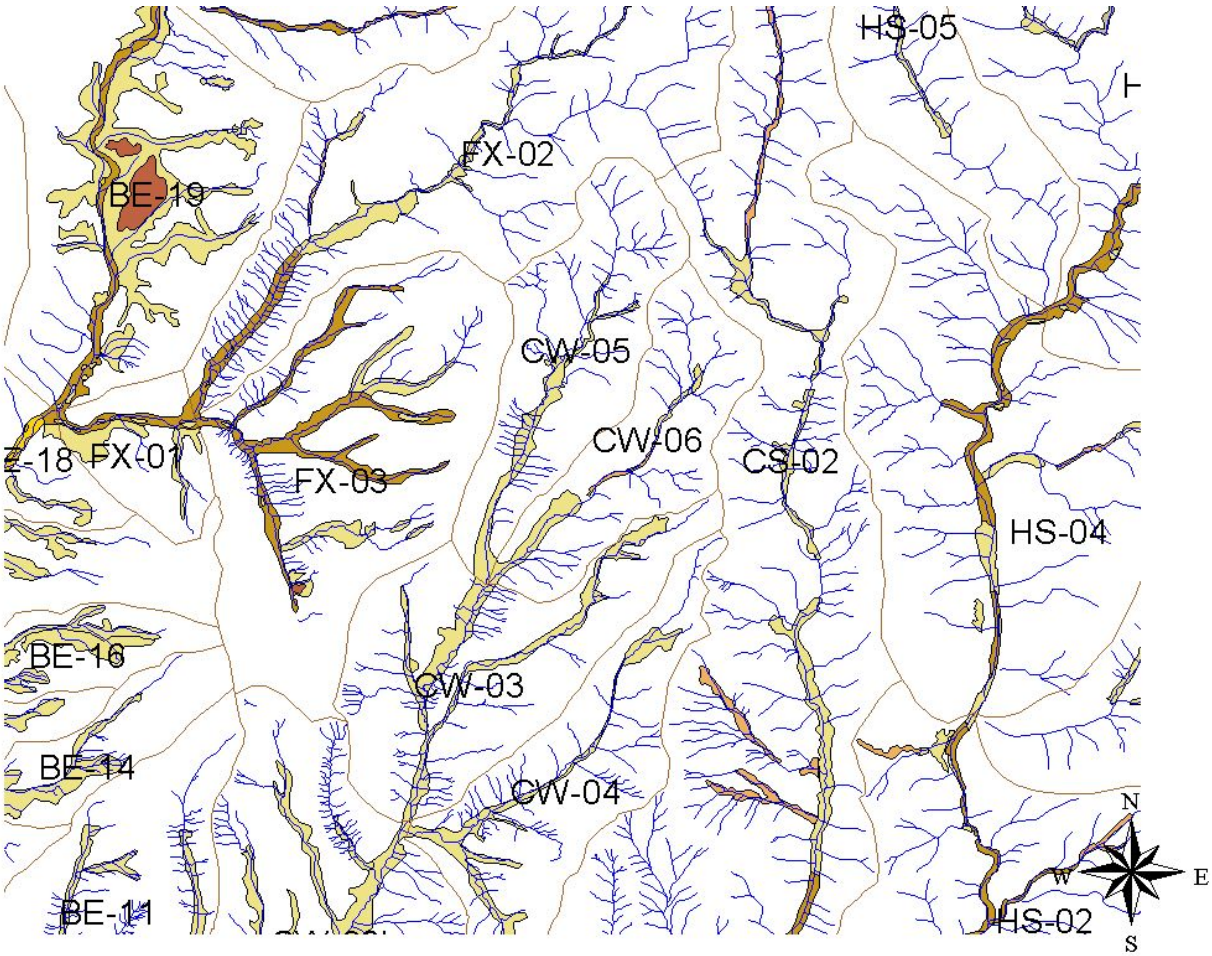


Figure 8. Riparian ecosystems in the Borrego and Serrano Creeks area consisting of stream channels (blue lines) and wetlands (colored polygons) delineated by Lichvar (2000)

Figure 9 is a bar chart showing indictor scores for each of the riparian reaches in the Borrego Canyon drainage basin. This type of bar chart can be quickly displayed in ArcView as a hotlinked image, in the spreadsheet and database as a hyperlink.

Figure 10 illustrates how the results can be used to summarize the baseline condition for riparian ecosystem integrity in the watershed, and compare riparian reaches. In the figure, labels represent the hydrologic integrity index (i.e., the sum of scores for indicators associated with hydrologic integrity). The highest possible score is 30 since six indicators are used to calculate the index. The gradient, from light to dark green, provides an easy way to visually identify areas of high hydrologic integrity in the watershed. Users with personal knowledge of the watershed

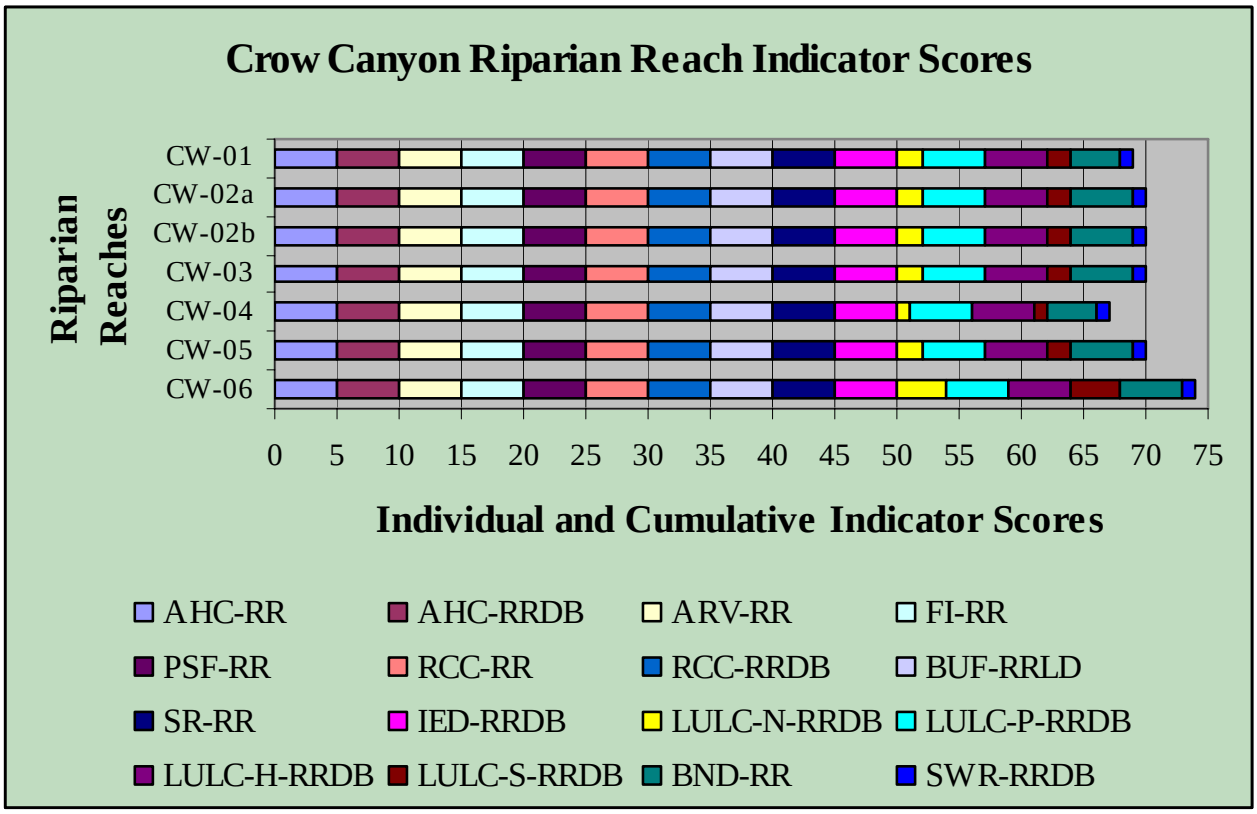


Figure 9. Scores for individual indicators in the Borrego Canyon drainage basin (indicator codes are provided in the Indicator Definitions, Metrics, and Reference Condition Section)

will recognize that index values are low in areas of urban and agricultural development and higher in the more remote, undeveloped areas. Figures 11and 12illustrate the results for water quality and habitat integrity indices with possible scores of 45 and 30 respectively.

The results also provide the mechanism for addressing the third task (i.e., to determine which of several proposed alternative development scenarios will result in the least impact to riparian ecosystem integrity in the watershed) identified in the Project Objectives above. By simulating the changes to indicators (i.e., changes in land use, hydraulic conveyance, etc.) that can be expected to occur as a result of a proposed alternative scenario, the existing information and tools can be used to generate new indictor scores and indices for riparian reaches. These scores and indices can then be compared with baseline indicator scores and integrity indices to show

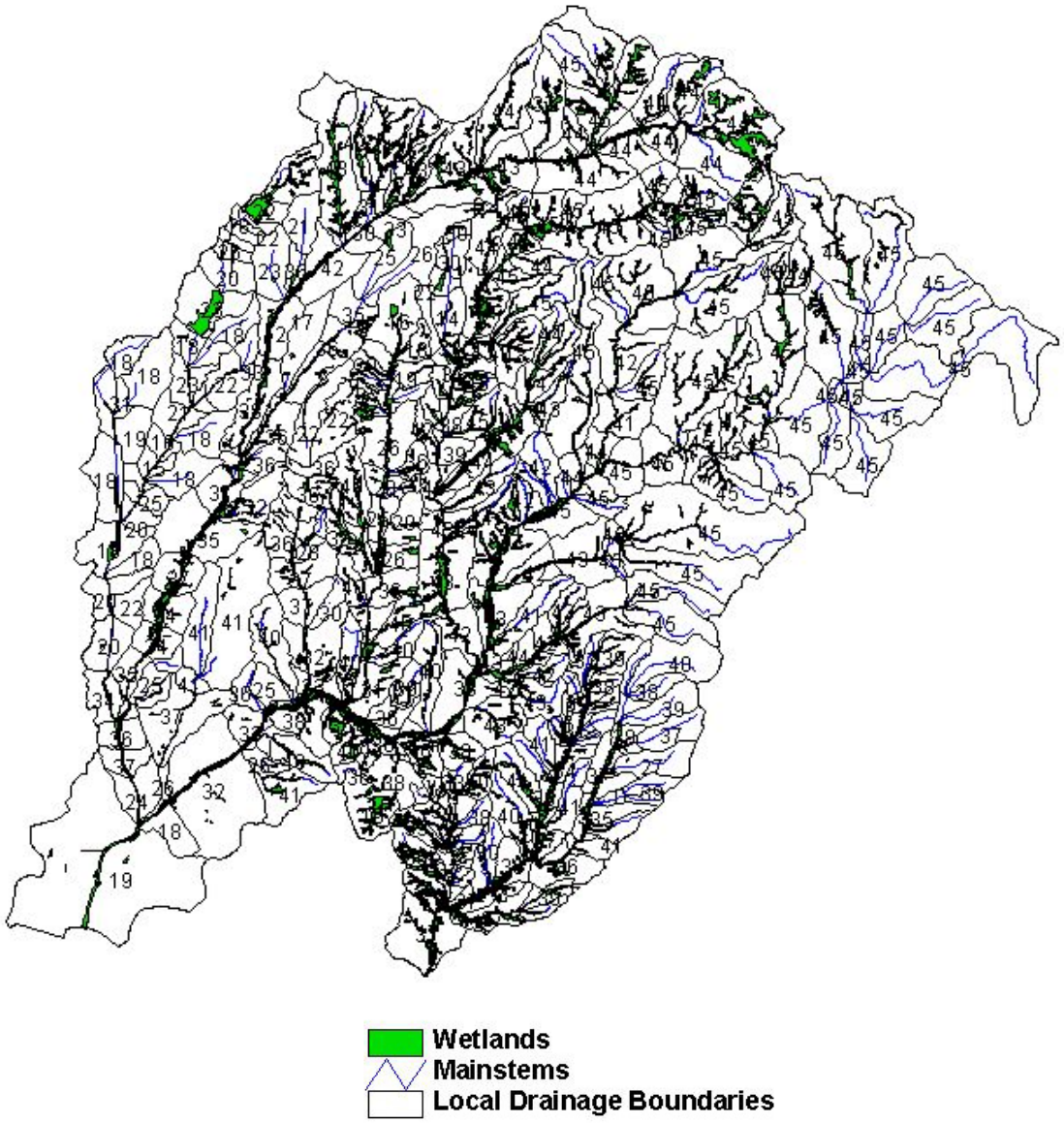


Figure 10 Hydrologic integrity indices (i.e., sum of hydrologic indicator scores with a possible total of 30) for riparian reaches in San Juan/San Mateo Watersheds. Indices are located in the local drainage area of the respective riparian reach.

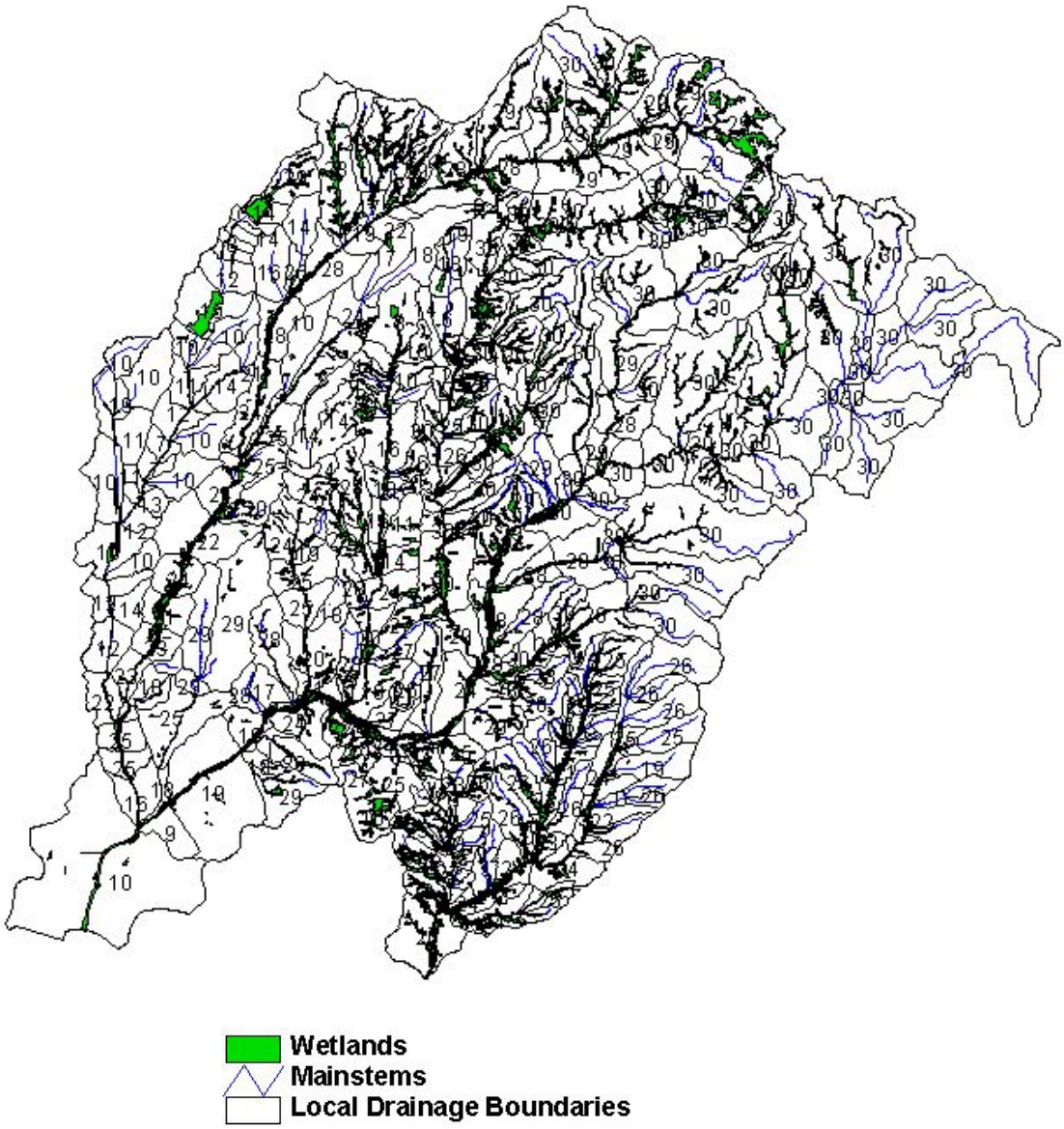


Figure 11 Water Quality integrity indices (i.e., sum of water quality indicator scores with a possible total of 45) for riparian reaches in San Juan/San Mateo Watersheds. Indices are located in the local drainage of the respective riparian reach.

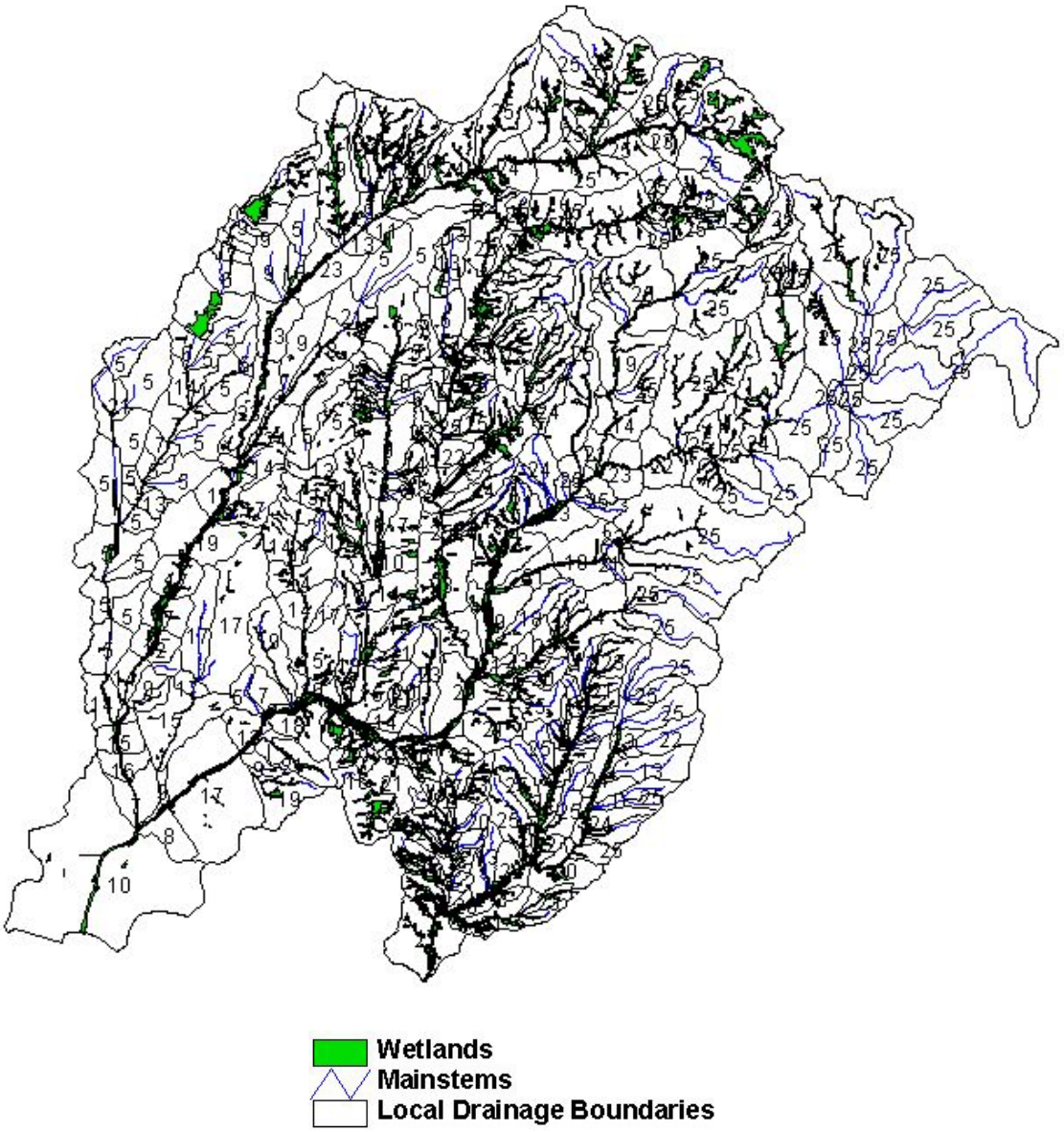


Figure 12 Habitat integrity indices (i.e., sum of habitat indicator scores with a possible total of 25) for riparian reaches in San Juan/San Mateo Watersheds. Indices are located in the local drainage area of the respective riparian reach.

how the proposed alternative scenarios will differentially impact riparian ecosystem integrity in the watershed.

We look forward to working with the Los Angeles District, other agencies, and stakeholders in completing this third task.

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* used as reference only, not cited in text

## 9.0 Appendix A: Spreadsheet Information and Guidelines

Information collected during field-work and information calculated during ArcView spatial analysis was input into a Microsoft Excel spreadsheet. Table 15 provides a listing of fields and date in this spreadsheet. This spreadsheet can be found in the following location on the CD: c:\san juan creek final\spreadsheets\san juan creekl.xls. The U. S. Army Corps of Engineers – Los Angeles District Regulatory Branch is responsible for distributing this information in an electronic format.

Table 15. List of spreadsheet fields and data for riparian reaches

| Riparian Reach Code | Drainage Basin | USGS 7.5 Minute Topograpic Quad | Main stem Downstream End Coordinates (UTM) | Main stem Upstream End Coordinates (UTM) | Size of Mapped Riparian Ecosystem in Riparian Reach Local Drainage (RRLD) (ha) |
|---|---|---|---|---|---|
| BE-01 | Borrega | El Toro | 11S 432113mE 3724084mN | 11S 432489mE 3724092mN | 0.6 |
| BE-02 | Borrega | El Toro | 11S 432489mE 3724092mN | 11S 434526mE 3725146mN | 2.3 |
| BE-03 | Borrega | El Toro | 11S 434526mE 3725146mN | 11S 435220mE 3725638mN | 1.9 |
| BE-04 | Borrega | El Toro | 11S 435220mE 3725638mN | 11S 435820mE 3725995mN | 3.7 |
| BE-05 | Borrega | El Toro | 11S 435820mE 3725995mN | 11S 436169mE 3726065mN | 2.5 |
| BE-06 | Borrega | El Toro | 11S 436169mE 3726065mN | 11S 436890mE 3726895mN | 1.1 |
| BE-07 | Borrega | El Toro | 11S 436169mE 3726065mN | 11S 437227mE 3726661mN | 0.0 |
| BE-08 | Borrega | El Toro | 11S 436890mE 3726895mN | 11S 437348mE 3727084mN | 0.8 |
| BE-09 | Borrega | El Toro | 11S 437361mE 3727089mN | 11S 438287mE 3727092mN | 0.0 |
| BE-10 | Borrega | El Toro | 11S 438287mE 3727092mN | 11S 439009mE 3728000mN | 0.0 |
| BE-11 | Borrega | El Toro | 11S 439009mE 3728000mN | 11S 440726mE 3729049mN | 0.0 |
| BE-12 | Borrega | El Toro | 11S 439009mE 3728000mN | 11S 439191mE 3729987mN | 0.0 |
| BE-13 | Borrega | El Toro | 11S 435239mE 3725741mN | 11S 435531mE 3726618mN | 0.5 |
| BE-14 | Borrega | El Toro | 11S 435531mE 3726618mN | 11S 437156mE 3727433mN | 1.2 |
| BE-15 | Borrega | El Toro | 11S 435531mE 3726618mN | 11S 436529mE 3727952mN | 0.6 |
| BE-16 | Borrega | El Toro | 11S 437159mE 3727601mN | 11S 437151mE 3728022mN | 0.0 |
| BE-17a | Borrega | El Toro | 11S 436804mE 3727941mN | 11S 437134mE 3728122mN | 0.0 |
| BE-17b | Borrega | El Toro | 11S 437482mE 3727771mN | 11S 438195mE 3728951mN | 0.0 |
| BE-18 | Borrega | El Toro | 11S 437134mE 3728122mN | 11S 438102mE 3729333mN | 0.0 |
| BE-19 | Bommer | Laguna Beach | 11S 426164mE 3722062mN | 11S 425230mE 3721572mN | 0.6 |
| BE-20 | Bommer | Laguna Beach | 11S 425230mE 3721572mN | 11S 425486mE 3721023mN | 1.1 |
| BE-21 | Bommer | Laguna Beach | 11S 424320mE 3720782mN | 11S 424251mE 3721545mN | 0.0 |
| BE-22 | Bommer | Laguna Beach | 11S 426097mE 3720505mN | 11S 425476mE 3721041mN | 0.4 |
| BE-23 | Bommer | Laguna Beach | 11S 424876mE 3720040mN | 11S 425349mE 3721177mN | 0.0 |
| BE-24 | Bommer | Laguna Beach | 11S 425625mE 3720006mN | 11S 425476mE 3721041mN | 0.9 |

Table 15. continued

| Riparian Reach Code | Size of Mapped Riparian Ecosystem in Riparian Reach Drainage Basin (RRDB) (ha) | Area of RRLD (ha) | Length of RRLD Perimeter (m) | Area of RRDB (ha) | Valley Type (Rosgen) | Valley Length (m) |
|---|---|---|---|---|---|---|
| BE-01 | 15.2 | 42 | 3687 | 1648 | X | 373 |
| BE-02 | 14.6 | 243 | 7053 | 1606 | X | 2304 |
| BE-03 | 12.2 | 130 | 5163 | 1363 | X | 848 |
| BE-04 | 8.0 | 98 | 5812 | 811 | X | 1124 |
| BE-05 | 4.3 | 15 | 1557 | 713 | X | 680 |
| BE-06 | 1.9 | 39 | 3058 | 573 | VII | 1631 |
| BE-07 | 0.0 | 124 | 5640 | 124 | VII | 1229 |
| BE-08 | 0.8 | 20 | 2228 | 535 | VII | 491 |
| BE-09 | 0.0 | 45 | 2741 | 514 | VII | 956 |
| BE-10 | 0.0 | 132 | 5413 | 469 | VII | 1262 |
| BE-11 | 0.0 | 208 | 7120 | 208 | II | 2288 |
| BE-12 | 0.0 | 129 | 5350 | 129 | II | 2135 |
| BE-13 | 2.3 | 37 | 2822 | 422 | X | 913 |
| BE-14 | 1.2 | 92 | 4918 | 228 | X | 1814 |
| BE-15 | 0.6 | 70 | 4625 | 157 | X | 1692 |
| BE-16 | 0.0 | 29 | 2428 | 137 | XI | 456 |
| BE-17a | 0.0 | 22 | 2158 | 88 | XI | 605 |
| BE-17b | 0.0 | 108 | 4629 | 108 | VII / II | 1541 |
| BE-18 | 0.0 | 66 | 4219 | 66 | VII / II | 1614 |
| BE-19 | 3.1 | 156 | 5994 | 657 | VII | 1045 |
| BE-20 | 0.9 | 50 | 3343 | 371 | VII | 624 |
| BE-21 | 3.9 | 44 | 2977 | 748 | VII / II | 1005 |
| BE-22 | 0.5 | 35 | 2834 | 85 | II | 834 |
| BE-23 | 0.0 | 49 | 3357 | 49 | II | 1269 |
| BE-24 | 0.9 | 63 | 4216 | 202 | VII | 1399 |

Table 15. continued

| Riparian Reach Code | Valley Width (m) | Main stem Downstream End Elevation (m) | Main stem Upstream End Elevation (m) | Valley Slope (%) (Estimated From 7.5 Minute Topo) | Engineered Channel Type or Rosgen Stream Type |
|---|---|---|---|---|---|
| BE-01 | no data | 74 | 84 | 2.7 | Engineered - Earthen and Riprap |
| BE-02 | no data | 93 | 122 | 1.3 | Engineered - Concrete |
| BE-03 | no data | 133 | 151 | 2.1 | D |
| BE-04 | no data | 151 | 169 | 1.6 | D |
| BE-05 | no data | 169 | 176 | 1.0 | D |
| BE-06 | 372 | 176 | 203 | 1.7 | G |
| BE-07 | 372 | 176 | 210 | 2.8 | Engineered - Earthen |
| BE-08 | 310 | 203 | 210 | 1.4 | Engineered - Concrete and Earthen |
| BE-09 | 320 | 212 | 243 | 3.2 | Engineered "Restored" |
| BE-10 | 110 | 243 | 281 | 3.0 | C and D |
| BE-11 | 30 | 281 | 422 | 6.2 | A and B |
| BE-12 | 25 | 281 | 452 | 8.0 | A and B |
| BE-13 | no data | 155 | 167 | 1.3 | C, D, and G |
| BE-14 | no data | 167 | 223 | 3.1 | B, C, and G |
| BE-15 | no data | 167 | 219 | 3.1 | B, C, and G |
| BE-16 | no data | 229 | 249 | 4.4 | Engineered - Earthen |
| BE-17a | no data | 230 | 248 | 3.0 | G and Engineered - Earthen |
| BE-17b | 45 | 245 | 441 | 12.7 | A and B |
| BE-18 | 44 | 258 | 430 | 10.7 | A and B |
| BE-19 | 375 | 68 | 86 | 1.7 | C and Engineered - Underground |
| BE-20 | 200 | 86 | 100 | 2.2 | C and B |
| BE-21 | 30 | 105 | 188 | 8.3 | A and B |
| BE-22 | 60 | 95 | 156 | 7.3 | B and G |
| BE-23 | 10 | 96 | 269 | 13.6 | A and B |
| BE-24 | 160 | 134 | 156 | 1.6 | B, C, and G |

Table 15. continued

| Riparian Reach Code | Main stem Channel Length in RRLD (m) (Smith) | Main stem Channel Length in RRDB (m) (Smith) | Main stem and Tributary Channel Length in RRLD (m) (Lichvar) | Main stem and Tributary Channel Length in RRDB (m) (Lichvar) | Main stem Channel Length / Main stem Channel and Tributary Channels Length | Drainage Density (Calculated) |
|---|---|---|---|---|---|---|
| BE-01 | 764 | 24844 | 395 | 63343 | 0.5 | 9 |
| BE-02 | 2418 | 24080 | 2766 | 62949 | 1.1 | 11 |
| BE-03 | 882 | 21662 | 1471 | 60182 | 1.7 | 11 |
| BE-04 | 805 | 11082 | 1990 | 38198 | 2.5 | 20 |
| BE-05 | 398 | 10277 | 395 | 36208 | 1.0 | 26 |
| BE-06 | 1118 | 9382 | 1131 | 34475 | 1.0 | 29 |
| BE-07 | 497 | 497 | 1338 | 1338 | 2.7 | 11 |
| BE-08 | 1283 | 8264 | 475 | 33343 | 0.4 | 23 |
| BE-09 | 1081 | 6981 | 780 | 32868 | 0.7 | 17 |
| BE-10 | 1342 | 5900 | 1533 | 32088 | 1.1 | 12 |
| BE-11 | 2468 | 2468 | 20168 | 20168 | 8.2 | 97 |
| BE-12 | 2090 | 2090 | 10386 | 10386 | 5.0 | 80 |
| BE-13 | 962 | 9698 | 948 | 20514 | 1.0 | 26 |
| BE-14 | 2426 | 4570 | 3171 | 12218 | 1.3 | 35 |
| BE-15 | 1927 | 4166 | 1916 | 7348 | 1.0 | 28 |
| BE-16 | 427 | 2144 | 641 | 9046 | 1.5 | 22 |
| BE-17a | 580 | 2239 | 621 | 5432 | 1.1 | 28 |
| BE-17b | 1717 | 1717 | 8405 | 8405 | 4.9 | 78 |
| BE-18 | 1659 | 1659 | 4811 | 4811 | 2.9 | 73 |
| BE-19 | 1279 | 10567 | 1392 | 27745 | 1.1 | 9 |
| BE-20 | 525 | 6682 | 2066 | 21061 | 3.9 | 41 |
| BE-21 | 1046 | 10865 | 1668 | 39215 | 1.6 | 38 |
| BE-22 | 1035 | 2434 | 1558 | 4409 | 1.5 | 45 |
| BE-23 | 1302 | 1302 | 3723 | 3723 | 2.9 | 76 |
| BE-24 | 1195 | 3504 | 2997 | 10307 | 2.5 | 47 |

Table 15. continued

| Riparian Reach Code | Channel Slope % (Calculated) | Sinuosity (Calculated) | Bankfull Width (ft) | Bankfull Width (m) | Floodprone Width (ft) | Floodprone Width (m) |
|---|---|---|---|---|---|---|
| BE-01 | 1.3 | 0.5 | 20.0 | 6.1 | 45.0 | 13.7 |
| BE-02 | 1.2 | 1.0 | no data | no data | no data | no data |
| BE-03 | 2.0 | 1.0 | 13.0 | 4.0 | 25.0 | 7.6 |
| BE-04 | 2.2 | 1.4 | no data | no data | no data | no data |
| BE-05 | 1.8 | 1.7 | 28.0 | 8.5 | 80.0 | 24.4 |
| BE-06 | 2.4 | 1.5 | 21.0 | 6.4 | 50.0 | 15.2 |
| BE-07 | 6.8 | 2.5 | no data | no data | no data | no data |
| BE-08 | 0.5 | 0.4 | no data | no data | no data | no data |
| BE-09 | 2.9 | 0.9 | no data | no data | no data | no data |
| BE-10 | 2.8 | 0.9 | 7.0 | 2.1 | 25.0 | 7.6 |
| BE-11 | 5.7 | 0.9 | 10.0 | 3.0 | 15.0 | 4.6 |
| BE-12 | 8.2 | 1.0 | 3.0 | 0.9 | 5.0 | 1.5 |
| BE-13 | 1.2 | 0.9 | 10.0 | 3.0 | 25.0 | 7.6 |
| BE-14 | 2.3 | 0.7 | 9.0 | 2.7 | 15.0 | 4.6 |
| BE-15 | 2.7 | 0.9 | 2.0 | 0.6 | 6.0 | 1.8 |
| BE-16 | 4.7 | 1.1 | no data | no data | no data | no data |
| BE-17a | 3.1 | 1.0 | no data | no data | no data | no data |
| BE-17b | 11.4 | 0.9 | 1.5 | 0.5 | 5.0 | 1.5 |
| BE-18 | 10.4 | 1.0 | 3.0 | 0.9 | 6.0 | 1.8 |
| BE-19 | 1.4 | 0.8 | no data | no data | no data | no data |
| BE-20 | 2.7 | 1.2 | 10.0 | 3.0 | 13.0 | 4.0 |
| BE-21 | 7.9 | 1.0 | 3.5 | 1.1 | 7.0 | 2.1 |
| BE-22 | 5.9 | 0.8 | 10.0 | 3.0 | 13.0 | 4.0 |
| BE-23 | 13.3 | 1.0 | 3.0 | 0.9 | 6.0 | 1.8 |
| BE-24 | 1.8 | 1.2 | 15.0 | 4.6 | 18.0 | 5.5 |

Table 15. continued

| Riparian Reach Code | Bankfull Maximum Depth (in) | Bankfull Maximum Depth (cm) | Bankfull Mean Depth (in) | Bankfull Mean Depth (cm) | Bankfull Cross-Sectional Area ($m^2$) | Width / Depth Ratio (Calculated) |
|---|---|---|---|---|---|---|
| BE-01 | 14.0 | 35.6 | 10.0 | 25.4 | 1.5 | 0.2 |
| BE-02 | no data | no data | no data | no data | no data | no data |
| BE-03 | 16.0 | 40.6 | 14.0 | 35.6 | 1.4 | 0.1 |
| BE-04 | no data | no data | no data | no data | no data | no data |
| BE-05 | 18.0 | 45.7 | 14.0 | 35.6 | 3.0 | 0.2 |
| BE-06 | 20.0 | 50.8 | 15.0 | 38.1 | 2.4 | 0.2 |
| BE-07 | no data | no data | no data | no data | no data | no data |
| BE-08 | no data | no data | no data | no data | no data | no data |
| BE-09 | no data | no data | no data | no data | no data | no data |
| BE-10 | 12.0 | 30.5 | 6.0 | 15.2 | 0.3 | 0.1 |
| BE-11 | 5.0 | 12.7 | 4.5 | 11.4 | 0.3 | 0.3 |
| BE-12 | 6.0 | 15.2 | 4.0 | 10.2 | 0.1 | 0.1 |
| BE-13 | 7.0 | 17.8 | 6.0 | 15.2 | 0.5 | 0.2 |
| BE-14 | 8.0 | 20.3 | 7.0 | 17.8 | 0.5 | 0.2 |
| BE-15 | 3.0 | 7.6 | 2.0 | 5.1 | 0.0 | 0.1 |
| BE-16 | no data | no data | no data | no data | no data | no data |
| BE-17a | no data | no data | no data | no data | no data | no data |
| BE-17b | 6.0 | 15.2 | 6.0 | 15.2 | 0.1 | 0.0 |
| BE-18 | 7.0 | 17.8 | 5.0 | 12.7 | 0.1 | 0.1 |
| BE-19 | no data | no data | no data | no data | no data | no data |
| BE-20 | 14.0 | 35.6 | 12.0 | 30.5 | 0.9 | 0.1 |
| BE-21 | 5.0 | 12.7 | 4.0 | 10.2 | 0.1 | 0.1 |
| BE-22 | 12.0 | 30.5 | 8.0 | 20.3 | 0.6 | 0.2 |
| BE-23 | 6.0 | 15.2 | 5.0 | 12.7 | 0.1 | 0.1 |
| BE-24 | 18.0 | 45.7 | 16.0 | 40.6 | 1.9 | 0.1 |

Table 15. continued

| Riparian Reach Code | Entrenchment Ratio (Calculated) | Natural Channel Substrate Bedrock / Boulder (%) | Natural Channel Substrate Cobble (%) | Natural Channel Substrate Gravel (%) | Natural Channel Substrate Sand (%) | Natural Channel Substrate Silt / Clay (%) |
|---|---|---|---|---|---|---|
| BE-01 | 2.3 | 0 | 5 | 10 | 80 | 5 |
| BE-02 | no data | no data | no data | no data | no data | no data |
| BE-03 | 1.9 | 0 | 5 | 10 | 80 | 5 |
| BE-04 | no data | 0 | 5 | 5 | 90 | 0 |
| BE-05 | 2.9 | 0 | 20 | 10 | 70 | 0 |
| BE-06 | 2.4 | 0 | 20 | 20 | 50 | 10 |
| BE-07 | no data | no data | no data | no data | no data | no data |
| BE-08 | no data | no data | no data | no data | no data | no data |
| BE-09 | no data | no data | no data | no data | no data | no data |
| BE-10 | 3.6 | 0 | 10 | 20 | 50 | 20 |
| BE-11 | 1.5 | 5 | 15 | 20 | 60 | 0 |
| BE-12 | 1.7 | 5 | 10 | 5 | 5 | 75 |
| BE-13 | 2.5 | 0 | 10 | 10 | 70 | 10 |
| BE-14 | 1.7 | 0 | 20 | 20 | 50 | 10 |
| BE-15 | 3.0 | 0 | 0 | 10 | 60 | 30 |
| BE-16 | no data | no data | no data | no data | no data | no data |
| BE-17a | no data | no data | no data | no data | no data | no data |
| BE-17b | 3.3 | 0 | 30 | 20 | 20 | 30 |
| BE-18 | 2.0 | 0 | 10 | 10 | 70 | 10 |
| BE-19 | no data | no data | no data | no data | no data | no data |
| BE-20 | 1.3 | 0 | 0 | 0 | 20 | 80 |
| BE-21 | 2.0 | 10 | 10 | 10 | 30 | 40 |
| BE-22 | 1.3 | 0 | 0 | 10 | 30 | 60 |
| BE-23 | 2.0 | 0 | 0 | 10 | 30 | 60 |
| BE-24 | 1.2 | 10 | 30 | 10 | 30 | 20 |

Table 15. continued

| Riparian Reach Code | Indicator 1 % of Channel in RRLD with Altered Hydraulic Conveyance (field observation) | Indicator 2 % of Channel in RRDB with Altered Hydraulic Conveyance (GIS) | Indicator 3 % of Area with Native Riparian Vegetation (field observation) | Indicator 4 % of Floodplain Present and not Isolated from Channel (field observation) | Indicator 5 % of Channel with Perennialized Stream Flow (field observation) |
|---|---|---|---|---|---|
| BE-01 | 100 | 36 | 50 | 50 | 100 |
| BE-02 | 100 | 34 | 0 | 0 | 100 |
| BE-03 | 80 | 27 | 80 | 60 | 0 |
| BE-04 | 10 | 37 | 100 | 100 | 0 |
| BE-05 | 10 | 39 | 100 | 100 | 0 |
| BE-06 | 100 | 37 | 50 | 25 | 0 |
| BE-07 | 100 | 100 | 0 | 0 | 100 |
| BE-08 | 100 | 29 | 0 | 100 | 100 |
| BE-09 | 100 | 15 | 0 | 0 | 0 |
| BE-10 | 0 | 0 | 100 | 100 | 100 |
| BE-11 | 0 | 0 | 100 | 100 | 0 |
| BE-12 | 0 | 0 | 100 | 100 | 100 |
| BE-13 | 0 | 10 | 75 | 50 | 0 |
| BE-14 | 0 | 9 | 100 | 50 | 0 |
| BE-15 | 0 | 14 | 100 | 100 | 0 |
| BE-16 | 100 | 20 | 0 | 0 | 0 |
| BE-17a | 100 | 26 | 50 | 25 | 0 |
| BE-17b | 0 | 0 | 100 | 100 | 0 |
| BE-18 | 0 | 0 | 100 | 100 | 0 |
| BE-19 | 50 | 6 | 25 | 50 | 50 |
| BE-20 | 0 | 0 | 90 | 100 | 0 |
| BE-21 | 0 | 23 | 25 | 100 | 0 |
| BE-22 | 0 | 0 | 100 | 100 | 0 |
| BE-23 | 0 | 0 | 80 | 80 | 0 |
| BE-24 | 0 | 0 | 100 | 100 | 0 |

Table 15. continued

| Riparian Reach Code | Indicator 6 % of Flood Prone Area in the RRLD with Riparian Corridor Breaks (field observation) | Indicator 7 % of Channel in the RRDB with Riparian Corridor Breaks (field observation) | Indicator 8 % of Channel in RRLD with Culturally Altered Buffer (100 m) (field observation) | Indicator 9 Sediment Regime Condition Index of Channel in RRLD (field observation) | Indicator 10 % of RRDB from which Surface Water is Imported, Exported or Diverted Water (GIS) | Indicator 11 % of LULC with the Potential to Contribute to a Nutrient Increase in Surface Waters (GIS) |
|---|---|---|---|---|---|---|
| BE-01 | 75 | 30 | 100 | 1 | 0 | 97 |
| BE-02 | 100 | 29 | 100 | 1 | 0 | 97 |
| BE-03 | 20 | 21 | 100 | 2 | 0 | 96 |
| BE-04 | 0 | 26 | 70 | 2 | 0 | 93 |
| BE-05 | 0 | 28 | 50 | 2 | 0 | 93 |
| BE-06 | 100 | 26 | 100 | 2 | 0 | 91 |
| BE-07 | 100 | 100 | 100 | 1 | 0 | 99 |
| BE-08 | 100 | 16 | 100 | 1 | 0 | 91 |
| BE-09 | 0 | 0 | 100 | 1 | 0 | 93 |
| BE-10 | 0 | 0 | 10 | 3 | 0 | 93 |
| BE-11 | 0 | 0 | 10 | 4 | 0 | 86 |
| BE-12 | 0 | 0 | 0 | 4 | 0 | 98 |
| BE-13 | 25 | 15 | 100 | 2 | 0 | 100 |
| BE-14 | 0 | 9 | 100 | 2 | 0 | 100 |
| BE-15 | 25 | 19 | 100 | 3 | 0 | 100 |
| BE-16 | 100 | 20 | 100 | 1 | 0 | 100 |
| BE-17a | 50 | 13 | 100 | 1 | 0 | 100 |
| BE-17b | 0 | 0 | 0 | 4 | 0 | 100 |
| BE-18 | 0 | 0 | 0 | 4 | 0 | 100 |
| BE-19 | 75 | 19 | 100 | 2 | 0 | 100 |
| BE-20 | 10 | 4 | 100 | 3 | 0 | 100 |
| BE-21 | 75 | 28 | 100 | 2 | 0 | 83 |
| BE-22 | 0 | 0 | 100 | 2 | 0 | 100 |
| BE-23 | 20 | 20 | 50 | 3 | 0 | 100 |
| BE-24 | 0 | 0 | 75 | 3 | 0 | 100 |

Table 15. continued

| Riparian Reach Code | Indicator 12 % of LULC with the Potential to Contribute to a Pesticide Increase in Surface Waters (GIS) | Indicator 13 % of LULC with the Potential to Contribute to a Hydrocarbon Increase in Surface Waters (GIS) | Indicator 14 % of LULC with the Potential to Contribute to a Sediment Increase in Surface Waters (GIS) | Indicator 15 % of RRLD with a Culturally Altered Boundary (GIS) | Indicator 16 % of RRDB affected by Surface Water Retention in Reservoirs and Detention Basins (GIS) |
|---|---|---|---|---|---|
| BE-01 | 97 | 17 | 90 | 100 | 0 |
| BE-02 | 97 | 16 | 90 | 100 | 0 |
| BE-03 | 96 | 11 | 96 | 100 | 0 |
| BE-04 | 93 | 9 | 93 | 100 | 0 |
| BE-05 | 93 | 7 | 93 | 100 | 0 |
| BE-06 | 91 | 7 | 91 | 72 | 0 |
| BE-07 | 99 | 0 | 99 | 0 | 0 |
| BE-08 | 91 | 4 | 91 | 100 | 0 |
| BE-09 | 93 | 4 | 93 | 5 | 0 |
| BE-10 | 93 | 5 | 93 | 0 | 0 |
| BE-11 | 86 | 11 | 86 | 0 | 0 |
| BE-12 | 98 | 0 | 98 | 0 | 0 |
| BE-13 | 100 | 16 | 100 | 100 | 0 |
| BE-14 | 100 | 25 | 100 | 69 | 0 |
| BE-15 | 100 | 2 | 100 | 31 | 0 |
| BE-16 | 100 | 0 | 100 | 100 | 0 |
| BE-17a | 100 | 0 | 100 | 0 | 0 |
| BE-17b | 100 | 0 | 100 | 0 | 0 |
| BE-18 | 100 | 0 | 100 | 0 | 0 |
| BE-19 | 100 | 4 | 96 | 63 | 0 |
| BE-20 | 100 | 0 | 100 | 100 | 0 |
| BE-21 | 83 | 6 | 83 | 100 | 0 |
| BE-22 | 100 | 0 | 100 | 100 | 0 |
| BE-23 | 100 | 0 | 100 | 100 | 0 |
| BE-24 | 100 | 0 | 100 | 100 | 0 |

Table 15. continued

| Riparian Reach Code | Indicator 1 Score % of Channel in RRLD with Altered Hydraulic Conveyance (field observation) | Indicator 2 Score % of Channel in RRDB with Altered Hydraulic Conveyance (GIS) | Indicator 3 Score % of Flood Prone Area with Riparian Vegetation (field observation) | Indicator 4 Score % of Floodplain Present and not Isolated from Channel (field observation) | Indicator 5 Score % of Channel with Perennialized Stream Flow (field observation) |
|---|---|---|---|---|---|
| BE-01 | 1 | 2 | 2 | 2 | 1 |
| BE-02 | 1 | 2 | 1 | 1 | 1 |
| BE-03 | 1 | 3 | 3 | 2 | 5 |
| BE-04 | 4 | 2 | 5 | 5 | 5 |
| BE-05 | 4 | 2 | 5 | 5 | 5 |
| BE-06 | 1 | 2 | 2 | 1 | 5 |
| BE-07 | 1 | 1 | 1 | 1 | 1 |
| BE-08 | 1 | 3 | 1 | 5 | 1 |
| BE-09 | 1 | 3 | 1 | 1 | 5 |
| BE-10 | 5 | 5 | 5 | 5 | 1 |
| BE-11 | 5 | 5 | 5 | 5 | 5 |
| BE-12 | 5 | 5 | 5 | 5 | 1 |
| BE-13 | 5 | 4 | 3 | 2 | 5 |
| BE-14 | 5 | 4 | 5 | 2 | 5 |
| BE-15 | 5 | 4 | 5 | 5 | 5 |
| BE-16 | 1 | 3 | 1 | 1 | 5 |
| BE-17a | 1 | 3 | 2 | 1 | 5 |
| BE-17b | 5 | 5 | 5 | 5 | 5 |
| BE-18 | 5 | 5 | 5 | 5 | 5 |
| BE-19 | 2 | 4 | 1 | 2 | 2 |
| BE-20 | 5 | 5 | 4 | 5 | 5 |
| BE-21 | 5 | 3 | 1 | 5 | 5 |
| BE-22 | 5 | 5 | 5 | 5 | 5 |
| BE-23 | 5 | 5 | 3 | 3 | 5 |
| BE-24 | 5 | 5 | 5 | 5 | 5 |

Table 15. continued

| Riparian Reach Code | Indicator 6 Score % of Flood Prone Area in the RRLD with Riparian Corridor Breaks (field observation) | Indicator 7 Score % of Channel in the RRDB with Riparian Corridor Breaks (field observation) | Indicator 8 Score % of Channel in RRLD with Culturally Altered Buffer (100 m) (field observation) | Indicator 9 Score Sediment Regime Condition Index of Channel in RRLD (field observation) | Indicator 10 Score % of RRDB from which Surface Water is Imported, Exported or Diverted Water (GIS) | Indicator 11 Score % of LULC with the Potential to Contribute to a Nutrient Increase in Surface Waters (GIS) |
|---|---|---|---|---|---|---|
| BE-01 | 1 | 2 | 1 | 1 | 5 | 1 |
| BE-02 | 1 | 3 | 1 | 1 | 5 | 1 |
| BE-03 | 3 | 3 | 1 | 2 | 5 | 1 |
| BE-04 | 5 | 3 | 1 | 2 | 5 | 1 |
| BE-05 | 5 | 3 | 2 | 2 | 5 | 1 |
| BE-06 | 1 | 3 | 1 | 2 | 5 | 1 |
| BE-07 | 1 | 1 | 1 | 1 | 5 | 1 |
| BE-08 | 1 | 3 | 1 | 1 | 5 | 1 |
| BE-09 | 5 | 5 | 1 | 1 | 5 | 1 |
| BE-10 | 5 | 5 | 4 | 3 | 5 | 1 |
| BE-11 | 5 | 5 | 4 | 4 | 5 | 1 |
| BE-12 | 5 | 5 | 5 | 4 | 5 | 1 |
| BE-13 | 3 | 4 | 1 | 2 | 5 | 1 |
| BE-14 | 5 | 4 | 1 | 2 | 5 | 1 |
| BE-15 | 3 | 3 | 1 | 3 | 5 | 1 |
| BE-16 | 1 | 3 | 1 | 1 | 5 | 1 |
| BE-17a | 2 | 4 | 1 | 1 | 5 | 1 |
| BE-17b | 5 | 5 | 5 | 4 | 5 | 1 |
| BE-18 | 5 | 5 | 5 | 4 | 5 | 1 |
| BE-19 | 1 | 3 | 1 | 2 | 5 | 1 |
| BE-20 | 4 | 5 | 1 | 3 | 5 | 1 |
| BE-21 | 1 | 3 | 1 | 2 | 5 | 1 |
| BE-22 | 5 | 5 | 1 | 2 | 5 | 1 |
| BE-23 | 3 | 3 | 2 | 3 | 5 | 1 |
| BE-24 | 5 | 5 | 1 | 3 | 5 | 1 |

Table 15. continued

| Riparian Reach Code | Indicator 12 Score % of LULC with the Potential to Contribute to a Pesticide Increase in Surface Waters (GIS) | Indicator 13 Score % of LULC with the Potential to Contribute to a Hydrocarbon Increase in Surface Waters (GIS) | Indicator 14 Score % of LULC with the Potential to Contribute to a Sediment Increase in Surface Waters (GIS) | Indicator 15 Score % of RRLD with a Culturally Altered Boundary (GIS) | Indicator 16 Score % of RRDB affected by Surface Water Retention in Reservoirs and Detention Basins (GIS) |
|---|---|---|---|---|---|
| BE-01 | 1 | 3 | 1 | 1 | 1 |
| BE-02 | 1 | 3 | 1 | 1 | 1 |
| BE-03 | 1 | 4 | 1 | 1 | 1 |
| BE-04 | 1 | 4 | 1 | 1 | 1 |
| BE-05 | 1 | 4 | 1 | 1 | 1 |
| BE-06 | 1 | 4 | 1 | 1 | 1 |
| BE-07 | 1 | 5 | 1 | 5 | 1 |
| BE-08 | 1 | 5 | 1 | 1 | 1 |
| BE-09 | 1 | 5 | 1 | 5 | 1 |
| BE-10 | 1 | 5 | 1 | 5 | 1 |
| BE-11 | 1 | 4 | 1 | 5 | 1 |
| BE-12 | 1 | 5 | 1 | 5 | 1 |
| BE-13 | 1 | 3 | 1 | 1 | 1 |
| BE-14 | 1 | 3 | 1 | 1 | 1 |
| BE-15 | 1 | 5 | 1 | 2 | 1 |
| BE-16 | 1 | 5 | 1 | 1 | 1 |
| BE-17a | 1 | 5 | 1 | 5 | 1 |
| BE-17b | 1 | 5 | 1 | 5 | 1 |
| BE-18 | 1 | 5 | 1 | 5 | 1 |
| BE-19 | 1 | 5 | 1 | 1 | 1 |
| BE-20 | 1 | 5 | 1 | 1 | 1 |
| BE-21 | 1 | 4 | 1 | 1 | 1 |
| BE-22 | 1 | 5 | 1 | 1 | 1 |
| BE-23 | 1 | 5 | 1 | 1 | 1 |
| BE-24 | 1 | 5 | 1 | 1 | 1 |

Table 15. completed

| Riparian Reach Code | Sum of Hydrologic Indicator Scores (possible 30) | Sum of Water Quality Indicator Scores (possible 45) | Sum of Habitat Indicator Scores (possible 30) | Sum of all Indicator Scores (possible 80) |
|---|---|---|---|---|
| BE-01 | 8 | 17 | 7 | 26 |
| BE-02 | 7 | 15 | 7 | 25 |
| BE-03 | 14 | 24 | 11 | 37 |
| BE-04 | 19 | 31 | 15 | 46 |
| BE-05 | 19 | 31 | 16 | 47 |
| BE-06 | 12 | 21 | 8 | 32 |
| BE-07 | 6 | 14 | 9 | 28 |
| BE-08 | 12 | 20 | 7 | 32 |
| BE-09 | 12 | 20 | 17 | 42 |
| BE-10 | 20 | 32 | 24 | 57 |
| BE-11 | 25 | 37 | 24 | 61 |
| BE-12 | 21 | 33 | 25 | 59 |
| BE-13 | 19 | 29 | 12 | 42 |
| BE-14 | 19 | 31 | 16 | 46 |
| BE-15 | 23 | 35 | 14 | 50 |
| BE-16 | 12 | 20 | 7 | 32 |
| BE-17a | 12 | 21 | 14 | 39 |
| BE-17b | 25 | 37 | 25 | 63 |
| BE-18 | 25 | 37 | 25 | 63 |
| BE-19 | 13 | 21 | 7 | 33 |
| BE-20 | 24 | 35 | 15 | 52 |
| BE-21 | 21 | 29 | 7 | 40 |
| BE-22 | 23 | 35 | 17 | 53 |
| BE-23 | 22 | 32 | 12 | 47 |
| BE-24 | 24 | 36 | 17 | 54 |

## 10.0 Appendix B: GIS Information and Guidelines

Spatial information collected and utilized for spatial analysis during the project was collected and saved as an ArcView project file. All themes and images in the project file are in a UTM – NAD83 projection in meter units. The U. S. Army Corps of Engineers – Los Angeles District Regulatory Branch is responsible for distribution of this information.

The ArcView GIS information for the San Juan/San Mateo Watersheds was organized under the San Juan/San Mateo Watersheds Folder as follows.

**1. San Juan/San Mateo Watershed Folder (c:\san juan creek final)**

Project Folder (c:\san juan creek final/apr files/san juan creek.apr)

AVL File Folder (c:\san juan creek final\avl files\)

ArcView "avl" files for loading pre-selected settings for themes including:

geology.avl – "avl" settings for geology.shp
habitat indicator score sum.avl – "avl" setting for habitat indicator score sum.shp
hotlink.avl – "avl" settings for all hotlinked image shape files
hydro indicator score sum.avl – "avl" setting for hydro indicator score sum.shp
lulc.avl – "avl" settings for land use / land cover
roads.avl – "avl" settings for roads
rrld.avl – "avl" settings for ssurgo.shp
streams.avl – "avl" settings for ssurgo.shp
ssurgo.avl – "avl" settings for ssurgo.shp
wous1.avl – "avl" settings for wous1.shp with Lichvar (2000) ratings
wous2.avl – "avl" settings for wous2.shp with Lichvar (2000) ratings
wq indicator score sum.avl – "avl" setting for wq indicator score sum.shp

Image Files (c:\san juan creek final\image files\)

ArcView image files including:

usgstopo1.tif - USGS 7.5 minute topo coverager for most of watershed
usgstopo2.tif - USGS 7.5 minute topo coverager for southeast corner of watershed

eagle.tif – true color partial aerial photo coverage of southern portion of watershed

Shape Files (c:\san juan creek final\shape files nad83 zone 11 m)

ArcView theme shape files including:

habitat indicator score sum.shp – Habitat integrity indices for all riparian reaches

hydro indicator score sum.shp – Hydrologic integrity indices all riparian reaches

geology.shp: Surficial geology

lulc.shp - Land Use / Land Cover

roads.shp - Major roads in the watershed

rrld.shp - Riparian Reach Local Drainage Boundaries

streams.shp – Working version of stream network for San juan Creek (D. Smith)

surgo.shp - SURGO soils

watershed.shp – Watershed boundary

wous1.shp - Waters of the US vectors (R. Lichvar)

wous2.shp - Waters of the US polygons (R. Lichvar

wq indicator score sum.shp – Water quality integrity indices all riparian reaches

ArcView hotlinked images attached to shape files including:

aerial1.shp – Hotlink to scanned aerial photo 1 of riparian reach

aerial2.shp – Hotlink to scanned aerial photo 2 of riparian reach

aerial3.shp – Hotlink to scanned aerial photo 3 of riparian reach

aerial4.shp – Hotlink to scanned aerial photo 4 of riparian reach

aerial5.shp – Hotlink to scanned aerial photo 5 of riparian reach

aerial6.shp – Hotlink to scanned aerial photo 6 of riparian reach

aerial7.shp – Hotlink to scanned aerial photo 7 of riparian reach

channel view.shp – Hotlink to channel view photo

field sheet 1.shp – Hotlink to field data sheet 1 scanned image

field sheet 2.shp – Hotlink to field data sheet 2 scanned image

indicator scores graph.shp – Hotlink to graph of indicator scores for riparian reach

terrace overview.shp – Hotlink to terrace overview photo

valley overview.shp – Hotlink to valley overview photo

“JPG” Files (c:\san juan creek final\aerials), (c:\san juan creek final\hyperlinked pics), (c:\san juan creek final\graphs), and (c:\san juan creek final\field forms)

Files with a “jpg” extension. These include aerial photographs, valley, terrace, and channel photos of each riparian reach, data sheets, and indicator score graphs. These files are hyperlinked in the Excel spreadsheet discussed in Appendix A, and “hot linked” in ArcView. However, in order for the hot links to work in ArcView, you must purchase extension software to access “jpg” image files in ArcView because the hotlink procedure in the ArcView program limited to one link, and “jpg” image files are not supported. There are several ArcView extensions available. We have found the PowerLink extension software to perform well, but others might be just as suitable. PowerLink can be purchased on line from http://www.spatial-online.com/dev/overview.asp. for $70.

Project sponsors and other end users have at least two choices for using the information in the folders above. The easiest thing to do is to copy the san juan creek final folder along with the appropriate subfolders (i.e., apr files, avl files, image files, shape files nad83 zone11 m, aerials, field forms, hyperlinked pics, and graphs) to the c:\ drive of your computer and simply open the san juan creek project in the apr files folder in ArcView. If this is the option you choose, it is critical that the folder is copied to the c:\ drive, and that you do not change the names of any folders or files.

The other option is to build a new ArcView project using the files in the themes and images folders and your knowledge of ArcView.

## 11.0 Appendix C: Database Information and Guidelines

Selected information collected during field-work and calculated during ArcView spatial analysis was placed in a Microsoft Access database. The primary data access screens in this database are described below. The U. S. Army Corps of Engineers – Los Angeles District Regulatory Branch is responsible for distributing this information.

**Screen 1**

Drop down pick list of all riparian reaches

**Screen 2**

Drop down picklist or list of bullets for each of the following:

- Summary Report for Riparian Reach (see Figure 6 above).
- Valley Overview Photo
- Terrace Overview Photo
- Channel Photo
- Aerial Photo 1
- Aerial Photo 2
- Aerial Photo 3
- Aerial Photo 4
- Aerial Photo 5
- Aerial Photo 6
- Aerial Photo 7
- Field Data Sheet 1
- Field Data Sheet 2

Note that the aerial photos, pictures, and field data sheets in the database are the same as those available as hyperlinks in the Excel spreadsheet, and as hotlinks in ArcView.

## 12.0 Appendix D: Glossary

**Aquatic Resources**

All waters and water habitats including lakes, ponds, streams, rivers and adjoining riparian areas that they affect, marshes, vernal pools, seeps, flats, and other wetlands.

**Buffer (area, zone, or habitat)**

An intervening upland area or other form of barrier that separates wetlands from developed or disturbed areas and reduces the impacts on the wetlands that may result from human activities. The critical functions of a buffer (associated with an aquatic system) include shading, input of organic debris and coarse sediments, uptake of nutrients, stabilization of banks, interception of fine sediments, stormflow attenuation during high water events, protection from disturbance by humans and domestic animals, maintenance of wildlife habitat, and room for variation of aquatic system boundaries over time due to hydrologic or climate effects.

**Channel**

A natural stream or river, or an artificial feature such as a ditch or canal that exhibits features of bed and bank, and conveys water primarily unidirectional and downgradient.

**Channel Type**

Channel type refers to the Rosgen (1996) classification of streams which is based on channel slope, sinuosity, entrenchment, width to depth ratios, and channel substrate.

**Clean Water Act**

The federal law which establishes standards and procedures for limiting the discharge of fill and pollutants into waters of the United States.

**Compensatory Mitigation**

The restoration, creation, and preservation of wetlands and/or other aquatic resources for the purpose of compensating for adverse impacts on an aquatic resource caused by a permitted project or activity.

**Creation**

The conversion of a persistent non-aquatic site into an aquatic site. For the purpose of this plan, creation includes the conversion of sites that currently do not meet the definition of wetlands, even though these sites were wetlands prior to being permanently drained and/or covered by fill.

**Delineation**

A determination of the boundaries of a wetland or other aquatic site.

**Discharge**

The placement of dredged or fill material into waters of the United States, including wetlands, that may result in impacts to the aquatic system. Examples include the redeposition of material during excavation, mechanized land clearing, and ditching.

**Enhancement**

Improving existing functions of a low-quality or degraded aquatic resource or wetland.

**Ephemeral**

Ephemeral streams are defined as streams in which flow is attributable only to surface water runoff in response to precipitation.

**Fill Material**

Material taken from an upland site and used to change the bottom features of waters of the United States (includes soil, rock, vegetative material, debris, pilings).

**Floodplain**

The land adjacent to a stream or lake, built of alluvium and subject to repeated flooding.

**Functional Assessment**

The process by which the capacity of a wetland to perform a function is measured.

**Geomorphic**

A term referring to the shape of the land surface.

**General Permit**

A permit for a specific class of activities with minimal individual and cumulative impact within a specified area issued by the Corps, authorizing the discharge of dredged or fill material into waters of the United States, including wetlands.

**Habitat Conservation Plan (HCP)**

A program of the U.S. Fish and Wildlife Service (Federal) designed to offset any harmful effects that a proposed activity might have on federally-listed threatened and endangered species. The HCP process allows development to proceed while providing for the conservation of species and their habitat, and provisions for incidental take

**Hydrogeomorphic Wetland Class**

A method of categorizing wetlands based on their hydrologic and geomorphic characteristics. There are five basic hydrogeomorphic classes including riverine, depression, fringe, slope, and flat wetlands.

**In-kind Mitigation**

Mitigation that results in wetlands of the same habitat type that provide similar functions and values of the wetlands removed.

**Intermittent Stream**

Intermittent streams are defined as streams in which ground water maintained base flow occurs intermittently at different times of the year

**Individual Permit**

A Section 404 permit issued by the Corps for an individual project for which a specific review was conducted.

**Jurisdictional Wetlands**

Areas that meet the soil, vegetation, and hydrologic criteria described in the " Corps of Engineers Wetlands Delineation Manual" (Environmental Laboratory 1987).

**Mitigation**

Actions or project design features that reduce impacts on wetlands by avoiding, minimizing, or compensating for adverse effects.

**Mitigation Banking**

Use of a single site, suitable for wetlands enhancement, restoration, and/or creation, for the mitigation of impacts on wetlands that result from more than one project at other sites.

**Mitigation Sequencing**

Provisions in the EPA Section 404(b)(1) Guidelines (40 CFR 230.10) and the 1990 Corps/EPA MOA requiring avoidance and minimization of adverse impacts on the aquatic environment before compensatory mitigation may be considered.

**Nationwide Permit**

General permit issued nationally for specific types of activities resulting in discharges of dredge or fill material into waters of the United States with minimal individual and cumulative impacts. Nationwide permits have been issued for several categories of activities including wetland restoration projects, maintenance of existing facilities, road crossings, bank protection, and fills 3 acres or less in size in the headwaters of watersheds.

**Natural Community Conservation Plan (NCCP)**

A program of the Department of Fish and Game (State) that takes a broad-based ecosystem approach to planning for the protection and perpetuation of biological diversity. The NCCP process identifies and provides for the regional or areawide protection of plants, animals, and their habitats, while allowing compatible and appropriate economic activity. The primary objective of the NCCP program is the conservation of natural communities at the ecosystem scale while accommodating compatible land uses.

**Perennial Stream**

Perennial streams are defined as streams in which base flow is maintained year round by groundwater

**Reference Wetland**

A wetland or one of a group of wetlands within a relatively homogeneous biogeographical region that represents the variation that occurs within the type because of natural or human-induced causes.

**Restoration**

Actions taken which result in the re-establishment of aquatic site structure, processes, and functions in areas where the aquatic site has been altered, degraded, or destroyed.

**Stream Type**

Stream type refers to the Rosgen (1996) classification of streams which is based on channel slope, sinuosity, entrenchment, width to depth ratios, and channel substrate.

**Section 404 Permit**

The permit issued by the Corps under Section 404 of the Clean Water Act for authorizing the discharge of dredged or fill material into waters of the United States, including wetlands; also known as Corps permit, fill permit, Department of the Army permit, DA permit, individual permit, 404 permit.

**Section 404 (b)(1) Guidelines**

Substantive regulations promulgated at 40 CFR 230 in accordance with Section 404(b)(1) of the Clean Water Act that provide the standards for unacceptable adverse impacts on waters of the United States, including wetlands, used to determine whether a Section 404 permit should be issued. Generally, discharges of fill are allowed under the Guidelines only if no other environmentally less damaging practicable alternative is available, no significant degradation of the waters, no adverse impacts to threatened and endangered species, and if appropriate and practicable steps have been taken to sequentially avoid, minimize and mitigate adverse impacts on the aquatic ecosystem.

**Valley Type**

Valley type refers to the Rosgen (1996) classification of valleys which is based on valley slope, width, and shape.

**Waters of the United States (WoUS)**

Water bodies that are regulated under Section 404 of the Clean Water Act. It is the broadest category of regulated water bodies and includes wetlands along with nonwetland habitats, such as streams, rivers, lakes, ponds, bays, and oceans.

**Watershed**

A geographical area that drains to a major water body such as a river, lake, or creek, which is usually the water body for which the basin is named.

**Wetland**

Areas inundated or saturated by surface or ground water at a frequency and duration sufficient to support, and that under normal circumstances do support, a prevalence of vegetation typically adapted for life in saturated soil conditions. Wetlands generally include swamps, marshes, bogs, and similar areas.